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Historical Catch Estimate Reconstruction for the Pacific Ocean based on Shark Fin Trade Data (1980-2016)

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## Shelley Clarke ${ }^{1}$

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#### Abstract

This paper describes a methodology to estimate silky shark catches in the Eastern and Western Central Pacific Ocean by all fleets based on a characterization of the global shark fin trade. Catch estimates using this method have been used in ICCAT blue and shortfin mako assessments, IOTC blue shark assessments, and the previous WCPFC assessment of silky sharks. Estimates were constructed using four steps. First, estimates of the number and biomass of silky shark represented in the global shark fin trade in 2000 were reconstructed using triangular distributions in a WinBUGS model. These estimates were then adjusted using annual imports into Hong Kong for 1980-2016. Figures were then further adjusted based on the diminishing share of Hong Kong's shark fin trade as compared to the total global trade in recent years. Finally, these adjusted annual global estimates were scaled in a number of ways (by ocean area ( $\mathrm{km}^{2}$ ), by tuna and tuna-like species catch, and by longline effort) to represent potential shark catches in the Eastern and Western Central Pacific Ocean. It is important to note that these estimates capture only a portion of the potential silky shark catches (i.e. only those sharks' whose fins are internationally traded).


## 1 Introduction

Under the Common Oceans (Areas Beyond National Jurisdiction (ABNJ)) Tuna Project, the Western and Central Pacific Fisheries Commission (WCPFC), with support from the United Nations Food and Agriculture Organization (FAO), is executing a programme of shark and bycatch work. One of the components of this work involves conducting Pacific-wide shark stock assessments for the bigeye thresher shark, the southern hemisphere porbeagle, the silky shark and the whale shark. The silky shark assessment aims to update a previous WCPFC assessment (Rice \& Harley 2013, which used data through 2009) as well as to expand it to a Pacific-wide scale as called for by the Common Oceans (ABNJ) Tuna Project design.

The silky shark (Carcharhinus falciformis) has been identified by both WCPFC and the InterAmerican Tropical Tuna Commission (IATTC) as being depleted and in need of management, and was recently listed on the Convention on International Trade in Endangered Species (CITES) Appendix II. This study thus provides an opportunity to update information useful to fisheries managers in the WCPFC and IATTC, both of which currently have some form of no-retention conservation and management measure (CMM) for this species. Through collaborating to analyze data from both the Western and Central Pacific Ocean (WCPO) and Eastern Pacific Ocean (EPO), the study has the potential to elucidate basin-wide patterns for this highly migratory stock.

Lack of historical catch data is often a major obstacle to the assessment of shark species, and was a limiting factor in the IATTC's recent attempt to assess the silky shark in the EPO (IATTC 2014). This paper adapts and applies a methodology previously used to produce estimates of catches of sharks utilised in the shark fin trade for the International Commission for the Conservation of Atlantic Tunas (Clarke 2008, 2016), the Indian Ocean Tuna Commission (Clarke 2015), and the Western and Central Pacific Fisheries Commission (Clarke 2009). These estimates are not direct substitutes for species-specific catch time series primarily because they capture only a portion of the potential shark mortality, i.e. only those sharks' whose fins are internationally traded. As a result, figures produced by this study should be considered minimum estimates of shark mortality in the Pacific Ocean. Nevertheless, they may be useful for comparison with other, more conventional sources of catch data or as minimum plausible estimates if other catch series are not available.

## 2 Materials and Methods

## $2.1 \quad$ Data Sources

The algorithm for estimating the Pacific Ocean shark catch represented in historical shark fin trade data is based on Clarke (2008, 2009). It consists of four data components, each of which is discussed separately below:

1. Estimates, by species, of the number and biomass of sharks used in the global shark fin trade based on market sampling in 2000 (the "anchor point" estimates);
2. A standardized estimate of the quantity of shark fins imported to Hong Kong for each year of interest before and after 2000 based on customs statistics;
3. An estimate of the Hong Kong market share, relative to the global market, for each year of interest before and after 2000 based on expert judgment; and
4. Estimates of the proportion of the global total of shark fins that are derived from the Pacific Ocean (calculated using several alternative methods).

### 2.1.1 Component 1

The "anchor point" estimates of the number and biomass of sharks used in the global shark fin trade are taken from Clarke et al. (2006a). That study used matches of Chinese trade names and taxa from market sampling and genetic testing (Clarke et al. 2006b), in combination with 18 months of Hong Kong auction records to impute missing data and produce an annual estimate of traded fin weights by species and fin size category. These fin weights were then converted to number of sharks and biomass using a series of conversion factors. For each species, three independent estimates based on dorsal, pectoral and caudal fins, respectively, were produced and extrapolated using trade data to represent the global market. A composite estimate for all fin types was then produced using a mixture distribution computed with the density function for each fin position weighted proportional to its precision. Since a probabilistic modelling framework was applied, the results were presented as probability intervals.

Of the eleven categories of species, or groups of species, presented in that study, this analysis uses the results for silky sharks only. In number, the quantity of silky sharks utilized in the shark fin trade in 2000 was estimated at 0.795 million ( $95 \%$ probability interval of $0.368-2.008$ million). In biomass, the quantity was $45,460 \mathrm{t}$ ( $95 \%$ probability interval of $29,400-74,050 \mathrm{t}$ ). These estimates are based on shark fin trade for 2000 when Hong Kong imported 6,788 t of fins and was estimated to control 44-59\% of the global market (Clarke 2004a, Clarke et al. 2006a)

### 2.1.2 Component 2

Standardized estimates of the quantity of shark fin imported by Hong Kong in each year since 1980 were prepared from unpublished Hong Kong government records (HKSARG 2017). Prior to 1998, Hong Kong recorded imports of shark fins in dried or frozen ("salted") categories without distinguishing between processed and unprocessed fins. In order to avoid double-counting fins returning to Hong Kong after processing in Mainland China, imports from the Mainland prior to 1998 were subtracted from total imports following methods used by TRAFFIC (1996). In 1998 Hong Kong established separate customs codes for dried and frozen (i.e. the latter listed as "salted" in commodity coding lists), processed and unprocessed fins. After 1998, only unprocessed dried
and frozen fins were included in the annual totals. All frozen fin weights were normalized for water content by multiplying by 0.25 (Clarke 2004a).

Although the data series continues through to the present, changes in the commodity coding scheme in 2012, in parallel with concomitant reports of a sharp drop in both market demand and price, suggest that Hong Kong import data after 2011 may not reflect trends in shark catches to the same extent as prior data (Dent \& Clarke 2015, Eriksson \& Clarke 2015). This problem forces a choice between using the reduced import figures and assuming they are accurate, or making an additional adjustment to compensate for the potential biases since 2011. One way of imputing figures for 2012-2016 would be to assume that trade levels since 2011 vary in proportion to global levels of chondrichthyan capture production. This appears to be a reasonable assumption given the parallel trends in the Hong Kong import figures and the FAO-compiled global chondrichthyan capture production figures (Figure 1) ${ }^{1}$. Alternative values for Hong Kong shark fin imports for 2012-2016 were imputed using $H_{i}=H_{2011} \times \frac{C_{i}}{C_{2011}}$ where $H_{i}$ is the annual Hong Kong import value for years $i=2012$ to 2016, and $C_{i}$ is the FAO-compiled annual value for global capture production of chondrichthyans in year $i$ (Table 1).

### 2.1.3 Component 3

Hong Kong's share of the global shark fin trade was studied in detail for 1996-2000 and was calculated from empirical data to range from 44-59\% (Clarke et al. 2006a). Since reliable empirical data for estimating Hong Kong's market share in previous and subsequent years (i.e. 1980-1995 and 2001-2016) are lacking, ranges of values for these years were specified based on expert judgment.

Difficulties in estimating Hong Kong's share of the global trade in previous years (i.e. 1980-1995) are mainly due to the lack of access to customs statistics for other countries, especially Mainland China. Nevertheless, a general understanding of trade patterns in Hong Kong during the 1980s (Clarke et al. 2007) suggests that Hong Kong's market share was higher in 1980-1995 than during 1996-2000. The earliest accounts of the shark fin trade state that Hong Kong's share of world imports was 50\% (Tanaka 1994, based on data through 1990) or 85\% (Vannuccini 1999, based on 1992 data). A range of $65-80 \%$ was thus selected for the period 1980-1990. A transitional period for the shark fin trade in Hong Kong occurred in 1991-1995 as demand began to rise appreciably in Mainland China. It is likely that Hong Kong's share began to drop, but not to the extent observed in the period 1996-2000 (i.e. 44-59\%), thus a range of $50-65 \%$ was selected.

Estimation of Hong Kong's market share since 2000 is less plagued by data gaps but still subject to a number of potential biases. Previous analysis has shown that Hong Kong imports of shark fin rose at a rate of 6\% per year from 1992-2000 (Clarke 2004a), but afterwards stabilized with a slightly declining linear trend (Clarke et al. 2007). Hong Kong shark fin traders attribute this trend to a loss of market share to Mainland China. While this explanation is supported by the well-known liberalization of the Mainland China economy just prior to and as a result of entry to the World Trade Organization in December 2001, Mainland China’s shark fin imports do not show a strong trend of increase since 2000. One reason for this lack of trend may be that in 2000 Mainland China began importing frozen shark fins under a category previously used only for frozen shark meat and therefore from 2000 onward frozen fins, which comprise a substantial portion of the trade, are no

[^1]longer distinguishable in the statistics (Clarke 2004b). Complications in trade reporting by Mainland China and their implications for assessing global trade in shark fins are discussed in detail in Clarke et al. (2007). On balance it was considered that even without strong evidence of increasing imports by Mainland China, it was likely that Hong Kong's share of global trade declined sharply after 2000. A range of $30-50 \%$ was thus specified for 2001-2006 to account for the initial decline, and a lower range of $25-40 \%$ was specified for 2007-2016 as the trend is believed to have become even more pronounced including potential diversion to countries in Southeast Asia (Dent \& Clarke 2015).

### 2.1.4 Component 4

Three methods were used for proportioning global fin trade-based catch estimates to Pacific Oceanspecific quantities. Each of the resulting indices has its own inherent biases acting over the entire time series or over portions of the time series. Therefore, when patterns appear in results derived from one proportioning method only, careful consideration of the credibility of that particular proportioning method is warranted.

The first proportioning method is based on calculating the area of silky shark potential habitat in the Eastern and Western Central Pacific relative to its potential habitat in the world ocean as a whole. This method assumes that the silky shark is evenly distributed throughout global waters between the northern-most and southern-most extent of its range. For simplicity, this range was considered to be $30^{\circ} \mathrm{N}-30^{\circ} \mathrm{S}$ worldwide based on indicative ranges given in Compagno (1984). Using Google Earth tools the global ocean area between $30^{\circ} \mathrm{N}$ and $30^{\circ} \mathrm{S}$ was calculated as 189.295 million $\mathrm{km}^{2}$. The portion of this potential silky shark habitat lying within the Western and Central Pacific Fisheries Convention Area was 66.603 million $\mathrm{km}^{2}$ and the portion within the InterAmerican Tropical Tuna Convention Area was 46.410 million $\mathrm{km}^{2}$. The area of the two convention areas together between $30^{\circ} \mathrm{N}$ and $30^{\circ} \mathrm{S}$ is smaller than the sum of their individual areas due to an overlap area of 5.847 million $\mathrm{km}^{2}$. Therefore if taken together the WCPF and IATT Convention Areas between $30^{\circ} \mathrm{N}$ and $30^{\circ} \mathrm{S}$ extend over $107.166 \mathrm{~km}^{2}$. Based on these figures, the ratios of Pacific to global areas are

Eastern Pacific only: $\frac{46.410 \mathrm{M} \mathrm{km}^{2}}{189.295 \mathrm{M} \mathrm{km}^{2}}=0.245$
Western and Central Pacific only: $\frac{66.603 \mathrm{M} \mathrm{km}^{2}}{189.295 \mathrm{M} \mathrm{km}^{2}}=0.352$
Pacific: $\frac{107.166 \mathrm{M} \mathrm{km}^{2}}{189.295 \mathrm{M} \mathrm{km}^{2}}=0.566$
No plot is shown for this area-based proportioning method because the ratios are constant throughout the time series.

The second proportioning method involves scaling against a ratio of tuna and tuna-like species catches in global waters versus those in the Pacific Ocean. In previous applications of this methodology, the scale was defined using FAO capture production values for the group of target species defined as "tunas, bonitos and billfishes" in the FISHSTATJ system (FAO 2018). In this case, given that a portion of the Eastern Pacific tuna fleet shifts it's targeting in part of the year to dolphinfish (Coryphaena hippurus; Martínez-Ortiz et al. 2015), the catch of this species was also taken into account. In the FISHSTATJ system there were two possible categories representing this species: 'common dolphinfish' and 'dorado' ('mahi mahi' is not shown). Since the 'dorado' entries
were for countries fishing in the Atlantic, only 'common dolphinfish' was included. Dolphinfish never comprised more than $2 \%$ of the total reported catch of the target species group in the Western Pacific, but in the Eastern Pacific it comprised as much as $6-8 \%$ of the total in recent years reflecting the expansion trends described in Martínez-Ortiz et al. (2015). FISHSTATJ reports capture production figures for the Northeast, Southeast, Eastern Central, and Northwest, Southwest and Western Central Pacific making it easy to separate eastern and western areas without overlap. The total capture production figures for the target species group, and the resulting ratios, are given in Table 3. As shown in Figure 2, the ratios of WCPO to EPO target species catches are quite stable throughout the time series.

The third proportioning method involved constructing an index of fishing effort. Although a number of gear types catch silky sharks, longline fishing effort was considered the best index for proportioning effort on silky sharks by ocean. The main reason for this is that longline effort is most easily standardized across oceans on the basis of hooks fished. Furthermore Lawson (2011) and Rice (2012), which estimated both WCPO longline and purse seine catches, found that the WCPO longline fishery catches 2-4 times the quantity of silky sharks as the WCPO purse seine fishery. Although catches with purse seine gear are a significant component of the global catch of silky shark, catch rates are likely to be dependent on set type (e.g. unassociated (free school) versus various types of associated sets) and these catch rates by set type are not constant across oceans (Restrepo et al. 2017). Bearing in mind that this step is not estimating the global catch--it is only partitioning the global catch to each ocean--a longline effort index was considered the best option. One practical matter limiting this method is that are no 2016 longline effort figures available for the Atlantic or Eastern Pacific Oceans at this time. Therefore, the ratios for 2015 were used for 2016 as placeholders.

The number of longline hooks (in millions) fished annually in the Western and Central Pacific was obtained from a database of raised longline effort for the WCPO maintained by the Pacific Community (CES 2017). For the Eastern Pacific, nominal longline effort has only been published for fleets from China, Japan, Korea, French Polynesia, Chinese Taipei and the United States (IATTC 2017). Effort for other fleets (i.e. last column of Table A-9 in IATTC (2017)) was imputed using the average catch rates for all other reporting fleets for a given year and added to the published total effort. Longline effort in the Atlantic has been estimated under ICCAT's EFFDIS project through 2015 only (ICCAT 2018). Longline effort data for the Indian Ocean were provided in standardized units for 1980-2016 by the IOTC Secretariat (IOTC 2018). Each of these series, as well as the sum of global longline effort, and the ratio of Eastern, Western and total Pacific Ocean effort to global longline effort are shown in Table 4 and Figure 3.

### 2.2 Modelling Methods

The model was implemented with Markov chain Monte Carlo (MCMC) methods using the Gibbs sampler (Gelfand and Smith 1990) via OpenBUGS software version 3.2.3 rev 1012 (Imperial College London 2014). Since the original posterior distributions presented in Clarke et al. (2006a) require many hours of computing time to replicate, simplified representations of these complex distributions were approximated using triangular distributions (Step 1). Other uncertain parameters, such as Hong Kong's share of the global fin trade (Step 3), were specified as expert judgement-based ranges with uniformly distributed random variables. The annual quantity of Hong Kong imports (Step 2) and the proportioning indices (Step 4) were based on empirical data for each year, except for the geographic area which does not vary from year to year. Although there is uncertainty in these data it is not possible to quantify the variance and thus these parameters
were specified using deterministic equations. The model was executed in four steps covering each of the four data sources given above (Annex 1):

## Step 1

The probability distributions representing the range of estimates of silky shark in the global trade by number and biomass ( $0.795(0.368-2.008)$ million in number and $45.46(29.40-74.05) \mathrm{t}$ in biomass) were approximated as triangular distributions using the reported lower limit of the $95 \%$ probability interval as the minimum, the upper limit of the $95 \%$ probability interval as the maximum, and the median as the mode. The model drew a random variable from each of the triangular distributions representing each species' number or biomass in 2000 in each iteration.

## Step 2

Each random variable drawn in Step 1 was multiplied by the ratio of the standardized quantity of fins traded through Hong Kong in each year from 1980-1999 and 2001-2016 (Table 1) to the quantity of fins traded through Hong Kong in 2000 (i.e. 6,788 t). This step serves to scale the global species-specific number or biomass estimates from 2000 to quantities representing trade levels in each of the other years. Due to a lack of quantitative data on trends in species composition this step assumes that the proportion of silky shark in 2000 remains constant over the years 1980-2016. This appears to be a reasonable assumption based on a recent study of shark fin trimmings, a byproduct of the shark fin trade and a possible indicator of species composition (Fields et al. 2017). The Fields et al. (2017) study found that silky shark was the second-most common species in the market in 2014-2015 comprising 5.4\% (95\% confidence interval of 2.1-11.4\%) whereas Clarke et al. (2016b) in 2000 found silky shark to be the third-most common species comprising $3.5 \%$ ( $95 \%$ probability interval of 3.1-4.0\%) of the shark fin market.

## Step 3

Hong Kong's share in four alternative periods ( $S_{a}$ ), i.e. 1980-1990, 1991-1995, 2001-2006 and 2007-2016, relative to its share in 1996-2000 ( $0.44-0.59, S$ ) was specified as a series of uniformly distributed random variables using endpoints based on expert judgment (Section 2.1.3). The ratio of $S$ and $S_{a}$ was then computed and multiplied by the result from Step 2. The result of Step 3 is a species-specific number or biomass value representing sharks used in the global trade for each year from 1980-2016.

## Step 4

The final step required proportioning the annual values from Step 3 to the Eastern Pacific, Western Pacific and Pacific Ocean as a whole. Proportioning based on area used constants for the various regions over all years in the time series. The target species catch-based (Table 3 and Figure 2) and longline effort-based (Table 4 and Figure 3) proportioning methods applied unique values for each year as deterministic calculations.

The model was run for 10,000 iterations, and medians and $95 \%$ probability interval endpoints were sampled from the final 1,000 iterations.

## 3 Results

The algorithm outlined above will, by definition, produce the same patterns of results in number (Figures 4-6 and Annexes 2-4) and in biomass (Figure 7-9 and Annexes 5-7). This is because the same scaling factors were applied to the number and biomass anchor point estimates thus only the absolute value of the starting point for each metric differs. Furthermore, in two of the three cases (i.e. for all but the target species catch-based method) the scaling factors did not show a strong
trend (Figures 2 and 3) and so the trend from the Hong Kong fin imports (Figure 1; Step 2) and the proportion of trade believed to be passing through Hong Kong (Step 3) drove the overall result. In particular, the increase in the early 2000s reflects the sharp rise in Hong Kong imports, and the reduction in the Hong Kong trade thereafter is modulated by the declining share of Hong Kong in the global trade in recent years.

In the Western Pacific the area-based proportioning method, which is constant over time, produced the lowest estimates in most years (Figures 4 and 7). The target species catch-based method generally produced the highest estimates and these were slightly less than double the area-based method. These patterns suggest that the Western Pacific produces a greater share of the world's tuna catch than would be expected given its ocean area which seems reasonable given the productivity of WCPO waters. The effort-based method closely followed the area-based method until 2006 and thereafter produced gradually higher estimates which approximated those of the target catch-based method by the end of the time series (2016). This pattern reflects the trend of increasing longline effort in the Western Pacific relative to the global longline effort since 1998 (Figure 3). Probability intervals for the three Western Pacific estimation approaches largely overlap and range as high as double their medians in the later years of the time series.

The patterns in the Eastern Pacific estimates show similar trends to those in the Western Pacific but the absolute values of the different estimation approaches have a different rank order (Figures 5 and 8): the area-based proportioning method produced the highest estimates suggesting that the region's silky shark habitat is large relative to its proportionally smaller share of global tuna/dolphinfish catch and longline effort. This may be an actual reflection of fishing practices or may result from an inability to accurately quantify this region's catches and effort. Similar to the estimates for the Western Pacific, the probability intervals for the various estimates are as high as double their medians in recent years.

Estimates for the Pacific as a whole are less divergent between the three methods (Figures 6 and 9). This suggests that the share of Pacific sharks (relative to global) is more stable between methods than is the split between WCPO and EPO between methods. For the Pacific estimates, the target species catch-based method remains higher than the effort-based method. The low estimates in the west from the area-based method combine with the high estimates in the east from the area-based method to place the area-based estimates for the Pacific as a whole intermediate to the other two series. Probability intervals for Pacific-wide estimates are not quite symmetrical around the medians with lower bounds around $50 \%$ lower than the medians and upper bounds as high as 60$80 \%$ above the medians.

Focusing on the medians for the 2008-2016 period as a basis for comparison between regions, estimates in number ranged from 0.5-0.8 million silky sharks per year in the Western Pacific, to $0.15-0.35$ million silky sharks per year in the Eastern Pacific and 0.8-1.0 million silky sharks per year for the Pacific Ocean as a whole. In biomass the respective figures are 25,000-40,000 $t$ in the Western Pacific to 4,000 to $17,000 \mathrm{t}$ in the Eastern Pacific to $35,000-50,000 \mathrm{t}$ Pacific-wide.

## 4 Discussion

Catch data for most shark species are insufficient to support stock assessment, yet concerns about the status of shark populations continue to grow. Under such circumstances, development of alternative historic shark catch time series and careful evaluation of whether these alternative series can fill some of the existing critical data gaps is a worthwhile exercise.

The estimates produced by this study were based on "anchor point" estimates derived from a shark fin trade data set compiled in Hong Kong in 2000 (Clarke et al. 2006a). To date these are the only quantitative, species-specific estimates of the number of sharks represented in the shark fin trade and represent a snapshot of the center of the global shark fin trade at that time. Using these data to estimate the number and biomass of shark catches in the Pacific Ocean requires a number of assumptions, namely:

1. The species composition of the sampled portion of the Hong Kong shark fin trade in Clarke et al. (2006a) is representative of the species composition of shark catches in Pacific offshore fisheries. As discussed in Clarke et al. (2006b), there is a lack of information to evaluate the strength of this assumption, but there are no other datasets that are considered more representative.
2. The species composition of the fin trade observed in 2000, and the relationships between fin sizes/weights and whole shark weights observed at that time, are constant throughout the time series. While some stock composition shifting would be expected over time, there are few existing data with which to explore alternative assumptions. A recent study of the species composition of the fin trimmings (by-product) trade in Hong Kong (Fields et al. 2017) is not directly comparable but shows remarkable consistency with the earlier fin-based study's species composition findings, including with regard to silky sharks.
3. Silky shark fins found in the Hong Kong shark fin trade are equally likely to derive from the Pacific as from any other ocean. This appears to be a reasonable assumption given what is known regarding the distribution of this species and the global nature of the trade.

Overlying these assumptions is the fact that estimating catches based on shark fin trade data will necessarily underestimate the true quantities of sharks caught. First, the original "anchor point" estimates are in themselves conservative because they are based only on those fins which could be confirmed to derive from the species of interest. More than half ( $54 \%$ ) of the fins observed by Clarke et al. (2006a) could not be characterized by species and could have contained additional quantities of the species of interest (Clarke et al. 2006b). Second, only those sharks whose fins enter the international shark fin trade are enumerated. This is because there is no means in this study of accounting for mortality associated with sharks which are a) discarded dead with their fins attached; b) released with their fins attached but subsequently die due to injury or stress; or c) are retained but whose fins are either not used or used without being internationally traded. For these two reasons actual shark mortality is very likely to be greater than the estimates provided here.

Robust estimation requires use of a number of different algorithms to explore various assumptions and biases. However, this approach in combination with reporting of probability intervals rather than point estimates can lead to considerable uncertainty when drawing conclusions about the estimation results. It is thus important to discuss, qualitatively if necessary, the relative credibility of each of the five estimates (Figures 4-9 and Annexes 2-7).

Of the three proportioning methods (area, target species catch and longline effort), the most arbitrary is the area-based method. Setting catch proportional to geographic area makes the unlikely assumption that shark abundance and fishing operations are evenly distributed throughout the world between $30^{\circ} \mathrm{N}-30^{\circ} \mathrm{S}$. Therefore this method would only be preferred when both target species-based and effort-based indices are considered unreliable or unrepresentative.

The target species proportioning method is most credible when the catch of sharks is expected to be proportional to the catch of target species, and when the catch reporting for target species is reliable. In both the WCPO and EPO there is the potential for fisheries to be catching sharks while targeting species other than tunas, billfishes and dolphinfishes (including fisheries targeting sharks per se), and if this is common the target species catch-based method will not be accurate. In the EPO it is known that artisanal and industrial longline fisheries operating in coastal and offshore waters catch a considerable number of sharks while targeting a mixture of species (GonzalesPestana et al. 2014, Martínez et al. 2015, Siu \& Aires-da-Silva 2016). Furthermore, the target species catch by these gear types may be considerably underrepresented in the FISHSTATJ database due to the variable quality of fisheries statistics in the region (Siu \& Aires-da-Silva 2016). The target species catch-based method is not likely to provide a robust estimate for these fisheries. In the WCPO mixed species longline fisheries exist but probably do not comprise as large a portion of the total longline catch as in the EPO.

The third method was based on effort statistics, specifically longline effort in hooks. This method is usually preferred when shark catches are primarily taken by longline gear and when longline effort data are considered to be reliable. Even though Lawson (2011) estimated that longline fisheries catch three-fold (or more) the quantity of silky sharks caught by the purse seine fishery, silky sharks are the most common shark caught in purse seine fisheries and such catches are not negligible. Therefore, for silky sharks, the longline effort-based method would be biased by changes in the relative effort of longline versus purse seine gear over time, and is not recommended for either the WCPO or EPO for this reason. The effort-based method is expected to perform particularly poorly for the EPO because the index was derived from incomplete longline effort data holdings (IATTC 2017) and so would tend to deflate the catch estimates for that region. Potential biases due to different statistical procedures applied by each t-RFMO to standardize and raise effort figures are also a drawback of this method.

Taking all of these considerations into account, it is recommended to use the area-based proportioning method for the EPO (Figures 5 and 8). In this region the effort-based approach is weakened by incomplete longline effort statistics and the target species catch-based approach may be biased by a large proportion of vessels not targeting tunas, billfishes or dolphinfishes. The same issues are present in the WCPO, but to a lesser extent. In the WCPO, longline effort estimated for the WCPFC is known to be missing effort fished by fleets from Indonesia, the Philippines and Vietnam and this compromises the effort-based silky shark catch estimates. The tuna catch estimates held by the WCPFC would be similarly compromised, but it is considered that the FAO tuna catch statistics used in this paper would be incomplete for various countries around the world yet not disproportionately under-reporting WCPO catches. Therefore, the target species catchbased estimates are considered reasonable for the WCPO (Figures 4 and 7). If a combined catch series is required for the entire Pacific, it is recommended to use the area-based method (Figures 6 and 9), or a sum of the target species catch-based efforts for the WCPO and the area-based method for the EPO.

To assess the credibility of the estimates produced in this study it is important to compare them to existing estimates of silky shark catches derived from more traditional fishery-dependent sources. For the EPO, stock assessment work by IATTC for the silky shark in 2013 produced catch series for what were considered to be northern and southern stocks (Figure 10; IATTC 2014). The catch series for the northern stock, which comprises the majority of the catch, spans 1993-2010 and varies from a low of just over $10,000 t$ to a high of slightly more than $16,000 \mathrm{t}$. Catches from the southern stock annually add another $2,000 \mathrm{t}$ or less until 2003 and less than $1,000 \mathrm{t}$ annually
thereafter. In comparison to the preferred Eastern Pacific biomass trend estimated here (i.e. the area-based estimates; Figure 8), IATTC (2014) shows a decrease in catches from 1993-1998 which is not reflected in the trade-based series. However, both the magnitude and trend of the tradebased estimates post-1998 show a remarkable consistency with the IATTC (2014) series. Both datasets show an increase in the early 2000s to approximately $15,000 \mathrm{t}$, followed by a sharp decline to approximately $12,000 \mathrm{t}$ in 2006 and then a quick recovery and steady increase to approximately $17,000 \mathrm{t}$ at the end of the series.

For the WCPO, there are several existing silky shark catch series prepared by the Pacific Community (SPC; Lawson (2011), Rice (2012) and Peatman et al. $(2017,2018)$ ) from observer data (Table 5; Figure 11). It is important to note that these estimates are all based on the same dataset, i.e. observer data from the WCPFC Regional Observer Programme and national observer programmes held by SPC. The variation in the estimates for any given year should thus be based on the analytical methodology alone, not due to different data sources or sampling regimes. These estimates indicate the catches by the purse seine fleet are low relative to the longline fleet and generally at or below 100,000 silky sharks per year. The most recent estimates of Peatman et al. (2017) estimate the lowest values for the purse seine fishery. Rice (2012) and Peatman et al. (2018) estimate similar magnitudes of silky shark catches in the longline fishery and their total catches (i.e. longline + purse seine) are up to $150 \%$ those of Lawson (2011). All of these SPC estimates are considerably lower than the preferred WCPO catch series produced by this study (i.e. the target species-based estimates in Figure 4; Figure 11). This is not surprising as the SPC dataset is focused on the tropical Pacific east of the Philippines and so the SPC studies could not estimate other components of the fishery, in particular the potentially large silky shark catches in the Southeast Asia region. It is also important to note that purse seine observer coverage in the WCPO was low until 2010 (see Clarke 2017, Table 8) and that longline coverage remains below the required 5\% level for many fleets (Williams et al. 2017). All of the SPC combined purse seine and longline catch series as well as the trade-based series estimated in this study show sharply increasing trends in the 2000s with both the trade-based series and the Peatman et al. $(2017,2018)$ series peaking in 2012, followed by a sharp drop and a slight rebound.

This discussion highlights that while both trade-based and fishery-based catch estimation methods have merit, there are also some important uncertainties associated with both methods which cannot be resolved on the basis of existing information. Given the urgent need for improvement in historic catch data to support shark stock assessment, further study of these and other methods is strongly encouraged.

## 5 Acknowledgements

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Table 1. Number and biomass of silky sharks (median and 95\% probability interval) used in the global shark fin trade in 2000 (Clarke et al. 2006a).

| Shark Species | Number (million) | Biomass ('000 t) |
| :--- | :---: | :---: |
| Silky | $0.795(0.368-2.008)$ | $45.46(29.94-74.05)$ |

Table 2. Adjusted total imports of shark fin ( t ) to Hong Kong, 1980-2016 (HKSARG 2017; see text for adjustment methods) and Global Capture Production for ISSCAAP group "sharks, rays and chimaeras" (FAO 2018). The "anchor point" estimate is shown in bold. Due to changes in commodity codes and a sharp curtailment of trade beginning in 2012 (Dent \& Clarke 2015) import quantities recorded by Hong Kong (in red) were replaced with values imputed assuming the recent relationship between Hong Kong import quantities and FAO-reported Global Capture Production figures is maintained.

| Year | Hong <br> Kong <br> Shark Fin <br> Imports <br> (t) |  | Global Shark <br> Catches (t) | Year | Hong <br> Kong <br> Shark <br> Fin <br> Imports <br> $\mathbf{( t )}$ | Global Shark <br> Catches (t) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1980 | 2,739 | 607,692 | 1999 | Imputed <br> Hong Kong <br> Shark Fin <br> Imports (t) |  |  |
| 1981 | 2,741 | 609,311 | $\mathbf{2 0 0 0}$ | $\mathbf{6 , 7 8 8}$ |  | 869,744 |
| 1982 | 2,704 | 614,179 | 2001 | 6,435 | 888,396 |  |
| 1983 | 2,512 | 562,173 | 2002 | 6,513 | 847,760 |  |
| 1984 | 2,748 | 602,724 | 2003 | 6,960 | 879,345 |  |
| 1985 | 2,613 | 624,012 | 2004 | 6,142 | 833,683 |  |
| 1986 | 2,788 | 627,731 | 2005 | 5,887 | 776,763 |  |
| 1987 | 3,317 | 661,790 | 2006 | 5,337 | 759,569 |  |
| 1988 | 3,272 | 688,459 | 2007 | 5,798 | 787,637 |  |
| 1989 | 3,003 | 679,241 | 2008 | 5,536 | 734,062 |  |
| 1990 | 3,018 | 701,752 | 2009 | 5,559 | 761,972 |  |
| 1991 | 3,526 | 723,551 | 2010 | 5,759 | 745,659 |  |
| 1992 | 4,265 | 740,549 | 2011 | 6,175 | 785,008 |  |
| 1993 | 3,856 | 753,331 | 2012 | 3,553 | 798,712 |  |
| 1994 | 4,144 | 770,017 | 2013 | 3,325 | 788,370 | $\mathbf{6 , 2 8 2}$ |
| 1995 | 4,706 | 776,535 | 2014 | 3,308 | 764,666 | $\mathbf{6 , 2 0 1}$ |
| 1996 | 4,513 | 826,809 | 2015 | 3,442 | 743,650 | $\mathbf{6 , 0 1 4}$ |
| 1997 | 4,868 | 852,323 | 2016 | 3,597 | 767,152 | $\mathbf{6 , 0 3 4}$ |
| 1998 | 5,196 | 843,100 |  |  |  |  |

Table 3. Estimates of FAO-reported capture production of tunas, bonitos, billfishes and dolphinfishes globally and in the Pacific Ocean as a whole, the EPO (Eastern Central, Northeast and Southeast) and the WCPO
(Western Central Pacific, Northwest, Southwest) 1980-2016 (FAO 2018). All catch values are in million $t$.

| Year | FAO <br> Global <br> Catch <br> Total | Pacific <br> Ocean <br> Catch <br> Total | Pacific:: <br> Global | Eastern <br> Pacific <br> Ocean <br> (EPO) <br> Catch <br> Total | Western <br> and <br> Central <br> Pacific <br> Ocean <br> (WCPO) | EPO: <br> Global | WCPO:: <br> Global |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1980 | 2.691 | 1.849 | 0.687 | 0.494 | 1.355 | 0.183 | 0.503 |
| 1981 | 2.724 | 1.808 | 0.664 | 0.515 | 1.293 | 0.189 | 0.475 |
| 1982 | 2.845 | 1.769 | 0.622 | 0.433 | 1.336 | 0.152 | 0.470 |
| 1983 | 3.003 | 1.950 | 0.649 | 0.371 | 1.579 | 0.123 | 0.526 |
| 1984 | 3.174 | 2.116 | 0.667 | 0.432 | 1.684 | 0.136 | 0.531 |
| 1985 | 3.275 | 2.068 | 0.631 | 0.589 | 1.479 | 0.180 | 0.452 |
| 1986 | 3.584 | 2.341 | 0.653 | 0.582 | 1.759 | 0.162 | 0.491 |
| 1987 | 3.729 | 2.419 | 0.649 | 0.594 | 1.825 | 0.159 | 0.489 |
| 1988 | 4.145 | 2.662 | 0.642 | 0.667 | 1.995 | 0.161 | 0.481 |
| 1989 | 4.164 | 2.703 | 0.649 | 0.662 | 2.041 | 0.159 | 0.490 |
| 1990 | 4.434 | 2.958 | 0.667 | 0.676 | 2.282 | 0.152 | 0.515 |
| 1991 | 4.584 | 3.072 | 0.670 | 0.589 | 2.483 | 0.129 | 0.542 |
| 1992 | 4.607 | 2.996 | 0.650 | 0.677 | 2.319 | 0.147 | 0.503 |
| 1993 | 4.707 | 2.849 | 0.605 | 0.602 | 2.247 | 0.128 | 0.477 |
| 1994 | 4.840 | 2.999 | 0.619 | 0.613 | 2.386 | 0.127 | 0.493 |
| 1995 | 4.993 | 3.093 | 0.619 | 0.665 | 2.428 | 0.133 | 0.486 |
| 1996 | 4.939 | 3.004 | 0.608 | 0.639 | 2.365 | 0.129 | 0.479 |
| 1997 | 5.218 | 3.295 | 0.631 | 0.761 | 2.533 | 0.146 | 0.485 |
| 1998 | 5.824 | 3.825 | 0.657 | 0.762 | 3.064 | 0.131 | 0.526 |
| 1999 | 5.977 | 3.874 | 0.648 | 0.867 | 3.007 | 0.145 | 0.503 |
| 2000 | 5.909 | 3.865 | 0.654 | 0.803 | 3.062 | 0.136 | 0.518 |
| 2001 | 5.862 | 3.851 | 0.657 | 0.889 | 2.962 | 0.152 | 0.505 |
| 2002 | 6.241 | 4.181 | 0.670 | 0.926 | 3.255 | 0.148 | 0.522 |
| 2003 | 6.392 | 4.239 | 0.663 | 1.035 | 3.204 | 0.162 | 0.501 |
| 2004 | 6.514 | 4.240 | 0.651 | 0.918 | 3.322 | 0.141 | 0.510 |
| 2005 | 6.660 | 4.293 | 0.645 | 0.916 | 3.377 | 0.138 | 0.507 |
| 2006 | 6.665 | 4.339 | 0.651 | 0.887 | 3.451 | 0.133 | 0.518 |
| 2007 | 6.734 | 4.638 | 0.689 | 0.781 | 3.857 | 0.116 | 0.573 |
| 2008 | 6.651 | 4.613 | 0.694 | 0.943 | 3.670 | 0.142 | 0.552 |
| 2009 | 6.769 | 4.648 | 0.687 | 0.979 | 3.670 | 0.145 | 0.542 |
| 2010 | 6.768 | 4.607 | 0.681 | 0.879 | 3.728 | 0.130 | 0.551 |
| 2011 | 6.696 | 4.480 | 0.669 | 0.961 | 3.519 | 0.144 | 0.526 |
| 2012 | 7.214 | 4.869 | 0.675 | 1.044 | 3.826 | 0.145 | 0.530 |
| 2013 | 7.345 | 4.945 | 0.673 | 1.042 | 3.904 | 0.142 | 0.531 |
| 2014 | 7.634 | 5.267 | 0.690 | 1.132 | 4.135 | 0.148 | 0.542 |
| 2015 | 7.457 | 5.123 | 0.687 | 1.228 | 3.895 | 0.165 | 0.522 |
| 2016 | 7.566 | 5.045 | 0.667 | 1.102 | 3.943 | 0.146 | 0.521 |

Table 4. Estimates of longline fishing effort (in million hooks) compiled from t-RFMO databases, and the ratio of the EPO, WCPO and Pacific effort to the global total, 1980-2016 (see text for derivation details).

| Year | Atlantic (ICCAT 2015) | Western <br> and <br> Central <br> Pacific <br> Longline <br> Effort <br> (CES <br> 2017) | Eastern <br> Pacific <br> Longline <br> Effort <br> (IATTC <br> 2017) | Indian <br> Ocean <br> Longline Effort <br> (IOTC <br> 2018) | Global <br> Total <br> Longline Effort | Ratio <br> (Western <br> Pacific <br> Ocean : <br> Global <br> Total) | Ratio (Eastern Pacific Ocean : Global Total) | Ratio (Pacific Ocean: Global Total) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 186 | 647 | 153 | 267 | 1253 | 0.517 | 0.122 | 0.639 |
| 1981 | 198 | 693 | 157 | 254 | 1301 | 0.532 | 0.121 | 0.653 |
| 1982 | 246 | 643 | 143 | 302 | 1334 | 0.482 | 0.107 | 0.589 |
| 1983 | 192 | 786 | 147 | 329 | 1454 | 0.541 | 0.101 | 0.642 |
| 1984 | 214 | 660 | 135 | 301 | 1310 | 0.504 | 0.103 | 0.607 |
| 1985 | 260 | 970 | 130 | 300 | 1660 | 0.584 | 0.078 | 0.662 |
| 1986 | 290 | 762 | 196 | 333 | 1581 | 0.482 | 0.124 | 0.606 |
| 1987 | 203 | 1043 | 238 | 361 | 1845 | 0.565 | 0.129 | 0.694 |
| 1988 | 284 | 908 | 236 | 416 | 1844 | 0.492 | 0.128 | 0.620 |
| 1989 | 287 | 817 | 230 | 529 | 1864 | 0.438 | 0.123 | 0.562 |
| 1990 | 292 | 884 | 238 | 568 | 1983 | 0.446 | 0.120 | 0.566 |
| 1991 | 294 | 752 | 284 | 573 | 1903 | 0.395 | 0.149 | 0.544 |
| 1992 | 312 | 866 | 271 | 649 | 2098 | 0.413 | 0.129 | 0.542 |
| 1993 | 326 | 687 | 225 | 856 | 2093 | 0.328 | 0.108 | 0.436 |
| 1994 | 385 | 556 | 224 | 815 | 1980 | 0.281 | 0.113 | 0.394 |
| 1995 | 397 | 570 | 191 | 738 | 1896 | 0.301 | 0.101 | 0.401 |
| 1996 | 380 | 539 | 153 | 916 | 1987 | 0.271 | 0.077 | 0.348 |
| 1997 | 334 | 537 | 141 | 974 | 1986 | 0.270 | 0.071 | 0.341 |
| 1998 | 326 | 611 | 177 | 1210 | 2325 | 0.263 | 0.076 | 0.339 |
| 1999 | 359 | 699 | 169 | 1043 | 2270 | 0.308 | 0.074 | 0.382 |
| 2000 | 351 | 751 | 149 | 937 | 2188 | 0.343 | 0.068 | 0.411 |
| 2001 | 324 | 934 | 274 | 869 | 2401 | 0.389 | 0.114 | 0.503 |
| 2002 | 325 | 974 | 345 | 863 | 2508 | 0.388 | 0.138 | 0.526 |
| 2003 | 352 | 962 | 338 | 785 | 2437 | 0.395 | 0.139 | 0.533 |
| 2004 | 370 | 1032 | 241 | 908 | 2550 | 0.405 | 0.094 | 0.499 |
| 2005 | 288 | 839 | 174 | 913 | 2214 | 0.379 | 0.079 | 0.458 |
| 2006 | 301 | 866 | 164 | 854 | 2185 | 0.396 | 0.075 | 0.472 |
| 2007 | 348 | 980 | 119 | 937 | 2385 | 0.411 | 0.050 | 0.461 |
| 2008 | 309 | 1001 | 110 | 723 | 2143 | 0.467 | 0.051 | 0.519 |
| 2009 | 328 | 1078 | 127 | 709 | 2243 | 0.481 | 0.057 | 0.538 |
| 2010 | 311 | 1047 | 156 | 650 | 2165 | 0.484 | 0.072 | 0.556 |
| 2011 | 326 | 1137 | 188 | 619 | 2270 | 0.501 | 0.083 | 0.584 |
| 2012 | 309 | 1191 | 191 | 680 | 2371 | 0.502 | 0.081 | 0.583 |
| 2013 | 272 | 979 | 221 | 705 | 2178 | 0.450 | 0.102 | 0.551 |
| 2014 | 212 | 1012 | 198 | 620 | 2042 | 0.495 | 0.097 | 0.593 |
| 2015 | 195 | 1078 | 220 | 525 | 2018 | 0.534 | 0.109 | 0.643 |
| 2016 | na | 1016 | na | 549 | 1564 | na | na | $n a$ |

Table 5. Summary of median silky shark catch estimates (in thousand sharks) for longline (LL) and purse seine (PS) fisheries in the WCPO from various sources (see Figure 11).

|  | $\begin{aligned} & \hline \text { Lawson } \\ & \text { (2011) } \end{aligned}$ | $\begin{aligned} & \hline \text { Lawson } \\ & (2011) \end{aligned}$ | $\begin{aligned} & \hline \text { Lawson } \\ & \text { (2011) } \end{aligned}$ | $\begin{aligned} & \hline \text { Rice } \\ & \text { (2012) } \end{aligned}$ | $\begin{aligned} & \hline \text { Rice } \\ & \text { (2012) } \end{aligned}$ | $\begin{aligned} & \hline \text { Rice } \\ & \text { (2012) } \end{aligned}$ | Peatman et al. (2017) | Peatman et al. <br> (2018) | Peatman et al. <br> (2017, <br> 2018) | Clarke (2018 (this paper from Figure 4 and Annex 3, target species catchbased method)) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | LL | PS | LL+PS | LL | PS | LL+PS | PS | LL | LL+PS | LL+PS |
| 1994 | 16,000 |  | 16,000 |  |  |  |  |  |  | 261,300 |
| 1995 | 161,000 | 23,800 | 184,800 | 271,970 | 34,480 | 306,450 |  |  |  | 292,500 |
| 1996 | 140,000 | 24,561 | 164,561 | 369,340 | 41,960 | 411,300 |  |  |  | 314,800 |
| 1997 | 135,000 | 28,102 | 163,102 | 118,510 | 73,160 | 191,670 |  |  |  | 343,800 |
| 1998 | 165,000 | 27,422 | 192,422 | 104,520 | 54,470 | 158,990 |  |  |  | 398,000 |
| 1999 | 167,000 | 35,172 | 202,172 | 237,160 | 59,520 | 296,680 |  |  |  | 426,600 |
| 2000 | 163,000 | 31,358 | 194,358 | 191,850 | 67,290 | 259,140 |  |  |  | 512,000 |
| 2001 | 149,000 | 35,069 | 184,069 | 241,920 | 50,870 | 292,790 |  |  |  | 609,100 |
| 2002 | 142,000 | 43,042 | 185,042 | 200,580 | 62,750 | 263,330 |  |  |  | 637,200 |
| 2003 | 97,000 | 56,544 | 153,544 | 183,570 | 96,100 | 279,670 | 42,951 | 238,945 | 281,896 | 653,500 |
| 2004 | 103,000 | 84,679 | 187,679 | 181,880 | 135,670 | 317,550 | 59,858 | 246,898 | 306,756 | 587,100 |
| 2005 | 114,000 | 78,976 | 192,976 | 134,380 | 83,840 | 218,220 | 55,283 | 238,827 | 294,110 | 559,400 |
| 2006 | 133,000 | 81,454 | 214,454 | 209,570 | 89,750 | 299,320 | 54,583 | 251,590 | 306,173 | 518,100 |
| 2007 | 167,000 | 78,999 | 245,999 | 338,400 | 88,990 | 427,390 | 51,385 | 318,992 | 370,377 | 766,600 |
| 2008 | 185,000 | 78,904 | 263,904 | 326,310 | 96,850 | 423,160 | 49,538 | 282,462 | 332,000 | 705,200 |
| 2009 | 189,000 | 69,790 | 258,790 | 389,520 | 99,090 | 488,610 | 42,830 | 403,173 | 446,003 | 695,300 |
| 2010 |  | 47,861 |  |  |  |  | 31,252 | 423,555 | 454,807 | 732,200 |
| 2011 |  |  |  |  |  |  | 51,947 | 435,988 | 487,935 | 749,500 |
| 2012 |  |  |  |  |  |  | 36,616 | 430,492 | 467,108 | 768,300 |
| 2013 |  |  |  |  |  |  | 41,476 | 191,521 | 232,997 | 759,800 |
| 2014 |  |  |  |  |  |  | 49,696 | 134,411 | 184,107 | 752,200 |
| 2015 |  |  |  |  |  |  | 40,323 | 177,214 | 217,537 | 704,500 |
| 2016 |  |  |  |  |  |  | 61,738 | 226,453 | 288,191 | 725,400 |



Figure 1. Annual imports of unprocessed shark fins, adjusted for water content, by Hong Kong 1980-2016 and global capture production of chondrichthyan fishes (sharks, rays and chimaeras) as reported to FAO 1980-2016.


Figure 2. Annual proportion of FAO-reported capture production of tunas, bonitos, billfishes and dolphinfish in the Pacific as a whole, the Eastern Pacific and Western Pacific as a proportion of the total global catch of these species, 1980-2016 (data given in Table 3).


Figure 3. Annual ratios of longline effort in the Eastern Pacific, Western Pacific Ocean and Pacific Ocean (as a whole) to global longline effort, 1980-2015. Data for 2016 are incomplete (see Table 4).


Figure 4. Annual median (solid line) and 95\% confidence interval (dashed lines) estimates for silky shark (in million sharks), using area, longline effort and target species catch proportioning methods to scale the number of sharks present in the global shark fin trade to those derived from the WCPO, 1980-2016.


Figure 5. Annual median (solid line) and $95 \%$ confidence interval (dashed lines) estimates for silky shark (in million sharks), using area, longline effort and target species catch proportioning methods to scale the number of sharks present in the global shark fin trade to those derived from the EPO, 1980-2016.


Figure 6. Annual median (solid line) and 95\% confidence interval (dashed lines) estimates for silky shark (in million sharks) using area, longline effort and target species catch proportioning methods to scale sharks present in the global shark fin trade to those derived from the Pacific Ocean as a whole, 1980-2016.


Figure 7. Annual median (solid line) and 95\% confidence interval (dashed lines) estimates for silky shark (in thousand t) using area, longline effort and target species catch proportioning methods to scale the sharks present in the global shark fin trade to those derived from the WCPO, 1980-2016


Figure 8. Annual median (solid line) and 95\% confidence interval (dashed lines) estimates for silky shark (in thousand t) using area, longline effort and target species catch proportioning methods to scale the sharks present in the global shark fin trade to those derived from the EPO, 1980-2016


Figure 9. Annual median (solid line) and $95 \%$ confidence interval (dashed lines) estimates for silky shark (in thousand $t$ ) using area, longline effort and target species catch proportioning methods to scale the sharks present in the global shark fin trade to those derived from the Pacific Ocean as a whole, 1980-2016.


re 11. Median estimates for silky shark catches (in number of sharks) for the Western and Central Pacific Ocean from various sources, 1980-2016 (see Table 5).

## Annex 1. WinBUGS code

model
\{
\#these are HK's assumed share of the global totals in each period shar8090~dunif( $0.65,0.80$ )
shar9195~dunif $(0.50,0.65)$
shar9600~dunif $(0.44,0.59)$
shar0006~dunif $(0.30,0.50)$
shar0716~dunif( $0.25,0.40$ )
for ( z in 1:11)\{
ratio[z] <- shar9600/shar8090
\}
for (z in 12:16) \{
ratio[z] <- shar9600/shar9195
\}
for (z in 17:21)\{ \#for 1996-2000 (this is the base period)
ratio[z] <-1
\}
for (z in 22:27) \{
\#2001-2006
ratio[z] <- shar9600/shar0006
\}
for (z in 28:37)\{
ratio[z] <- shar9600/shar0716 \#2007-2016
\}
\#for (g in 1:1) \{
\#this is a triangular distribution
\# $\quad$ rv[g]~dunif( 0,1000 )
\# $\mathrm{x}[\mathrm{g}]<-\mathrm{rv}[\mathrm{g}] / 1000$
\# gate[g]<-((trimode[g]-trimin [g]) / (trimax[g]-trimin[g]))
\# $\mathrm{A}[\mathrm{g}]<-\min (\mathrm{x}[\mathrm{g}], \mathrm{gate}[\mathrm{g}]) \quad$ \# find out whether x is higher or lower than criterion
\# $\quad \mathrm{B}[\mathrm{g}]<-$ equals $(\mathrm{x}[\mathrm{g}], \mathrm{A}[\mathrm{g}]) \quad$ \# if x IS lower then B will be 1 , if $\mathrm{x}>$ calculation then B will be 0
\# $\mathrm{C}[\mathrm{g}]<-$ equals $(\mathrm{B}[\mathrm{g}], 0) \quad$ \# sets C to zero if $\mathrm{B}=1$ or sets C to 1 if $\mathrm{b}=0$; so B and C are binary and opposite
\# $\quad \operatorname{draw}[\mathrm{g}]<-\left(\mathrm{B}[\mathrm{g}]^{*}\left(\operatorname{trimin}[\mathrm{~g}]+\operatorname{sqrt}\left(\mathrm{x}[\mathrm{g}] *(\operatorname{trimode}[\mathrm{~g}]-\operatorname{trimin}[\mathrm{g}])^{*}(\operatorname{trimax}[\mathrm{~g}]-\operatorname{trimin}[\mathrm{g}])\right)\right)\right)$
\# $\quad+\left(\mathrm{C}[\mathrm{g}]^{*}\left(\right.\right.$ trimax $\left.\left.[\mathrm{g}]-\mathrm{sqrt}\left((1-\mathrm{x}[\mathrm{g}])^{*}(\operatorname{trimax}[\mathrm{~g}]-\operatorname{trimode}[\mathrm{g}])^{*}(\operatorname{trimax}[\mathrm{~g}]-\operatorname{trimin}[\mathrm{g}])\right)\right)\right)$

```
\#this is a triangular distribution for 2000 (number ( N ) and then biomass ( B ))
```

for ( d in 1:2) \{
rv[d]~dunif( 0,1000 \#uninformative prior
$x[d]<-r v[d] / 1000$
gate[d]<-((trimode[d]-trimin[d]) / (trimax[d]-trimin[d]))
$A[d]<-\min (x[d], g a t e[d])$ \# find out whether $x$ is higher or lower than criterion $B[d]<-$ equals $(x[d], A[d]) \quad$ \# if $x$ IS lower then $B$ will be 1 , if $x>$ calculation then $B$ will be 0 $\mathrm{C}[\mathrm{d}]<-$ equals $(\mathrm{B}[\mathrm{d}], 0) \quad$ \# sets C to zero if $\mathrm{B}=1$ or sets C to 1 if $\mathrm{b}=0$; so B and C are binary and opposite draw $[\mathrm{d}]<-\left(\mathrm{B}[\mathrm{d}]^{*}\left(\operatorname{trimin}[\mathrm{~d}]+\operatorname{sqrt}\left(\mathrm{x}[\mathrm{d}]^{*}(\operatorname{trimode}[\mathrm{~d}]-\operatorname{trimin}[\mathrm{d}]) *(\operatorname{trimax}[\mathrm{~d}]-\operatorname{trimin}[\mathrm{d}])\right)\right)\right)$
$+\left(\mathrm{C}[\mathrm{d}]^{*}\left(\operatorname{trimax}[\mathrm{~d}]-\mathrm{sqrt}\left([1-\mathrm{x}[\mathrm{d}]) *(\operatorname{trimax}[\mathrm{~d}]-\operatorname{trimode}[\mathrm{d}])^{*}(\operatorname{trimax}[\mathrm{~d}]-\operatorname{trimin}[\mathrm{d}])\right)\right)\right)$
for (h in 1:37) \{
scaled[d,h] <- draw[d] * (HKimport[h]/HKimport[21])
share[d,h] <- scaled[d,h] * ratio[h] \#scale by whether HK's share was more or less than in 2000
areapropW[d,h] <- share[d,h] * GISW[d] \#area scalling for WCPO
areapropE[d,h] <- share[d,h] * GISE[d] \#area scaling for EPO
areaprop $[\mathrm{d}, \mathrm{h}]<-$ share[d,h]* GISP[d] \#area scaling for Pacific as a whole
tunapropW[d,h] <- share[d,h] * tunaW[h]
tunapropE[d,h] <- share[d,h] * tunaE[h]
tunapropP[d,h] <- share[d,h] *tunaP[h]
hookpropW[d,h] <- share[d,h] * LLW[h]
hookpropE[d,h] <- share[d,h] * LLE[h]
\#scale by total tuna catch (FISHSTAT) for WCPO
\#scale by total tuna catch for EPO
\#scale by LL hook effort for WCPO
\#scale by LL hook effort for EPO
\}
\}

## \#DATA

list(
\#NUMBER OF SHARKS (in millions) and BIOMASS (in '000 t)
trimin $=c(0.368,29.94)$,
trimode $=c(0.795,45.46)$,
trimax $=c(2.008,74.05)$,
HKimport=c(
2739,2741,2704,2512,2748, \#HK adjusted imports 1980-2016
2613,2788,3317,3272,3003, 3018,3526,4265,3856,4144, $4706,4513,4868,5196,5824$, 6788,6435,6513,6960,6142, 5887,5337,5798,5536,5559, 5759,6175,6282,6201,6014,5849,6034),

GISW=c(0.352,0.352),
GISE=c(0.245,0.245),
GISP $=c(0.566,0.566)$,
tunaW=c(
$0.503,0.475,0.470,0.526,0.531$, 0.452,0.491,0.489,0.481,0.490, $0.515,0.542,0.503,0.477,0.493$, $0.486,0.479,0.485,0.526,0.503$, $0.518,0.505,0.522,0.501,0.510$, $0.507,0.518,0.573,0.552,0.542$, $0.551,0.526,0.530,0.531,0.542$, $0.522,0.521$ ),
tunaE=c(
0.183,0.189,0.152,0.123,0.136, 0.180,0.162,0.159,0.161,0.159, 0.152,0.129,0.147,0.128,0.127, $0.133,0.129,0.146,0.131,0.145$, 0.136,0.152,0.148,0.162,0.141, $0.138,0.133,0.116,0.142,0.145$, $0.130,0.144,0.145,0.142,0.148$, $0.165,0.146$ ),
tunaP=c(
0.687,0.664,0.622,0.649,0.667, 0.631,0.653,0.649,0.642,0.649, $0.667,0.670,0.650,0.605,0.619$, $0.619,0.608,0.631,0.657,0.648$, 0.654,0.657,0.670,0.663,0.651, $0.645,0.651,0.689,0.694,0.687$, $0.681,0.669,0.675,0.673,0.690$, $0.687,0.667$ ),

LLP=c(
0.639,0.653,0.589,0.642,0.607, $0.662,0.606,0.694,0.620,0.562$, $0.566,0.544,0.542,0.436,0.394$, $0.401,0.348,0.341,0.339,0.382$, $0.411,0.503,0.526,0.533,0.499$, $0.458,0.472,0.461,0.519,0.538$, $0.556,0.584,0.583,0.551,0.593$, $0.643,0.643)$,

LLE=c(
0.122,0.121,0.107,0.101,0.103, 0.078,0.124,0.129,0.128,0.123, $0.120,0.149,0.129,0.108,0.113$, $0.101,0.077,0.071,0.076,0.074$,
$0.068,0.114,0.138,0.139,0.094$, 0.079,0.075,0.050,0.051,0.057, 0.072,0.083,0.081,0.102,0.097, $0.109,0.109)$,

LLW=c(
0.517,0.532,0.482,0.541,0.504, $0.584,0.482,0.565,0.492,0.438$, $0.446,0.395,0.413,0.328,0.281$, $0.301,0.271,0.270,0.263,0.308$, 0.343,0.389,0.388,0.395,0.405, 0.379,0.396,0.411,0.467,0.481, $0.484,0.501,0.502,0.450,0.495$, $0.534,0.534)$ )

Annex 2. Silky shark estimates in number (million sharks) for the area-based estimation method

|  | Eastern Pa mean | acific sd | MC error | 2.50\% | median | 97.50\% | Pacific mean | sd | MC error | 2.50\% | median | 97.50\% | Western P mean | Pacific <br> sd | MC error | 2.50\% | median | 97.50\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.07246 | 0.02465 | 7.76E-04 | 0.03394 | 0.06875 | 0.1272 | 0.1674 | 0.05696 | 0.001792 | 0.0784 | 0.1588 | 0.2938 | 0.1041 | 0.03542 | 0.001115 | 0.04876 | 0.09877 | 0.1827 |
| 1981 | 0.07251 | 0.02467 | $7.76 \mathrm{E}-04$ | 0.03396 | 0.0688 | 0.1273 | 0.1675 | 0.057 | 0.001794 | 0.07846 | 0.1589 | 0.294 | 0.1042 | 0.03545 | 0.001115 | 0.0488 | 0.09884 | 0.1828 |
| 1982 | 0.07154 | 0.02434 | $7.66 \mathrm{E}-04$ | 0.0335 | 0.06787 | 0.1255 | 0.1653 | 0.05623 | 0.001769 | 0.0774 | 0.1568 | 0.29 | 0.1028 | 0.03497 | 0.0011 | 0.04814 | 0.09751 | 0.1804 |
| 1983 | 0.06646 | 0.02261 | 7.12E-04 | 0.03113 | 0.06305 | 0.1166 | 0.1535 | 0.05223 | 0.001644 | 0.07191 | 0.1457 | 0.2694 | 0.09548 | 0.03249 | 0.001022 | 0.04472 | 0.09059 | 0.1676 |
| 1984 | 0.0727 | 0.02473 | $7.78 \mathrm{E}-04$ | 0.03405 | 0.06897 | 0.1276 | 0.168 | 0.05714 | 0.001798 | 0.07866 | 0.1593 | 0.2947 | 0.1045 | 0.03554 | 0.001118 | 0.04892 | 0.0991 | 0.1833 |
| 1985 | 0.06913 | 0.02352 | $7.40 \mathrm{E}-04$ | 0.03238 | 0.06558 | 0.1213 | 0.1597 | 0.05433 | 0.00171 | 0.0748 | 0.1515 | 0.2803 | 0.09932 | 0.03379 | 0.001063 | 0.04652 | 0.09423 | 0.1743 |
| 1986 | 0.07376 | 0.02509 | $7.90 \mathrm{E}-04$ | 0.03455 | 0.06998 | 0.1294 | 0.1704 | 0.05797 | 0.001824 | 0.07981 | 0.1617 | 0.299 | 0.106 | 0.03605 | 0.001135 | 0.04963 | 0.1005 | 0.186 |
| 1987 | 0.08775 | 0.02986 | $9.40 \mathrm{E}-04$ | 0.0411 | 0.08325 | 0.154 | 0.2027 | 0.06897 | 0.00217 | 0.09495 | 0.1923 | 0.3558 | 0.1261 | 0.0429 | 0.00135 | 0.05905 | 0.1196 | 0.2213 |
| 1988 | 0.08656 | 0.02945 | 9.27E-04 | 0.04054 | 0.08213 | 0.1519 | 0.2 | 0.06804 | 0.002141 | 0.09366 | 0.1897 | 0.3509 | 0.1244 | 0.04231 | 0.001332 | 0.05825 | 0.118 | 0.2183 |
| 1989 | 0.07945 | 0.02703 | $8.51 \mathrm{E}-04$ | 0.03721 | 0.07537 | 0.1394 | 0.1835 | 0.06244 | 0.001965 | 0.08596 | 0.1741 | 0.3221 | 0.1141 | 0.03883 | 0.001222 | 0.05346 | 0.1083 | 0.2003 |
| 1990 | 0.07984 | 0.02716 | $8.55 \mathrm{E}-04$ | 0.0374 | 0.07575 | 0.1401 | 0.1845 | 0.06276 | 0.001975 | 0.08639 | 0.175 | 0.3237 | 0.1147 | 0.03903 | 0.001228 | 0.05373 | 0.1088 | 0.2013 |
| 1991 | 0.1176 | 0.04006 | 0.00118 | 0.05454 | 0.1105 | 0.2001 | 0.2717 | 0.09255 | 0.002727 | 0.126 | 0.2552 | 0.4624 | 0.169 | 0.05756 | 0.001696 | 0.07836 | 0.1587 | 0.2876 |
| 1992 | 0.1423 | 0.04846 | 0.001428 | 0.06597 | 0.1336 | 0.2421 | 0.3286 | 0.112 | 0.003298 | 0.1524 | 0.3087 | 0.5593 | 0.2044 | 0.06962 | 0.002051 | 0.09479 | 0.192 | 0.3478 |
| 1993 | 0.1286 | 0.04381 | 0.001291 | 0.05965 | 0.1208 | 0.2189 | 0.2971 | 0.1012 | 0.002982 | 0.1378 | 0.2791 | 0.5057 | 0.1848 | 0.06295 | 0.001854 | 0.0857 | 0.1736 | 0.3145 |
| 1994 | 0.1382 | 0.04708 | 0.001387 | 0.0641 | 0.1298 | 0.2352 | 0.3193 | 0.1088 | 0.003205 | 0.1481 | 0.3 | 0.5434 | 0.1986 | 0.06765 | 0.001993 | 0.0921 | 0.1866 | 0.338 |
| 1995 | 0.157 | 0.05347 | 0.001575 | 0.07279 | 0.1475 | 0.2671 | 0.3626 | 0.1235 | 0.003639 | 0.1682 | 0.3407 | 0.6171 | 0.2255 | 0.07682 | 0.002263 | 0.1046 | 0.2119 | 0.3838 |
| 1996 | 0.1673 | 0.05407 | 0.001677 | 0.08004 | 0.161 | 0.2779 | 0.3864 | 0.1249 | 0.003874 | 0.1849 | 0.3719 | 0.642 | 0.2403 | 0.07768 | 0.002409 | 0.115 | 0.2313 | 0.3993 |
| 1997 | 0.1804 | 0.05832 | 0.001809 | 0.08633 | 0.1737 | 0.2998 | 0.4168 | 0.1347 | 0.004178 | 0.1994 | 0.4012 | 0.6925 | 0.2592 | 0.08379 | 0.002599 | 0.124 | 0.2495 | 0.4307 |
| 1998 | 0.1926 | 0.06225 | 0.001931 | 0.09215 | 0.1854 | 0.32 | 0.4449 | 0.1438 | 0.00446 | 0.2129 | 0.4282 | 0.7392 | 0.2767 | 0.08944 | 0.002774 | 0.1324 | 0.2663 | 0.4597 |
| 1999 | 0.2159 | 0.06978 | 0.002164 | 0.1033 | 0.2078 | 0.3586 | 0.4987 | 0.1612 | 0.004999 | 0.2386 | 0.48 | 0.8285 | 0.3101 | 0.1002 | 0.003109 | 0.1484 | 0.2985 | 0.5153 |
| 2000 | 0.2516 | 0.08132 | 0.002522 | 0.1204 | 0.2422 | 0.418 | 0.5812 | 0.1879 | 0.005826 | 0.2781 | 0.5594 | 0.9657 | 0.3615 | 0.1168 | 0.003623 | 0.173 | 0.3479 | 0.6006 |
| 2001 | 0.315 | 0.1171 | 0.003383 | 0.1387 | 0.2955 | 0.5837 | 0.7276 | 0.2706 | 0.007817 | 0.3205 | 0.6826 | 1.348 | 0.4525 | 0.1683 | 0.004861 | 0.1993 | 0.4245 | 0.8386 |
| 2002 | 0.3188 | 0.1185 | 0.003424 | 0.1404 | 0.2991 | 0.5907 | 0.7365 | 0.2739 | 0.007911 | 0.3243 | 0.6909 | 1.365 | 0.458 | 0.1703 | 0.00492 | 0.2017 | 0.4297 | 0.8487 |
| 2003 | 0.3407 | 0.1267 | 0.00366 | 0.15 | 0.3196 | 0.6313 | 0.787 | 0.2927 | 0.008454 | 0.3466 | 0.7383 | 1.458 | 0.4894 | 0.182 | 0.005258 | 0.2156 | 0.4592 | 0.907 |
| 2004 | 0.3006 | 0.1118 | 0.003229 | 0.1324 | 0.282 | 0.5571 | 0.6945 | 0.2583 | 0.007461 | 0.3059 | 0.6516 | 1.287 | 0.4319 | 0.1606 | 0.00464 | 0.1902 | 0.4052 | 0.8004 |
| 2005 | 0.2881 | 0.1072 | 0.003095 | 0.1269 | 0.2703 | 0.534 | 0.6657 | 0.2476 | 0.007151 | 0.2932 | 0.6245 | 1.234 | 0.414 | 0.154 | 0.004447 | 0.1823 | 0.3884 | 0.7672 |
| 2006 | 0.2612 | 0.09714 | 0.002806 | 0.115 | 0.2451 | 0.4841 | 0.6035 | 0.2244 | 0.006483 | 0.2658 | 0.5662 | 1.118 | 0.3753 | 0.1396 | 0.004032 | 0.1653 | 0.3521 | 0.6955 |
| 2007 | 0.3459 | 0.1247 | 0.003697 | 0.159 | 0.3278 | 0.6177 | 0.7992 | 0.2881 | 0.008542 | 0.3674 | 0.7573 | 1.427 | 0.497 | 0.1792 | 0.005312 | 0.2285 | 0.4709 | 0.8875 |
| 2008 | 0.3303 | 0.1191 | 0.00353 | 0.1518 | 0.313 | 0.5898 | 0.7631 | 0.2751 | 0.008156 | 0.3508 | 0.723 | 1.363 | 0.4746 | 0.1711 | 0.005072 | 0.2182 | 0.4497 | 0.8474 |
| 2009 | 0.3317 | 0.1196 | 0.003545 | 0.1525 | 0.3143 | 0.5923 | 0.7662 | 0.2762 | 0.008189 | 0.3523 | 0.726 | 1.368 | 0.4765 | 0.1718 | 0.005093 | 0.2191 | 0.4515 | 0.8509 |
| 2010 | 0.3436 | 0.1239 | 0.003672 | 0.158 | 0.3256 | 0.6136 | 0.7938 | 0.2861 | 0.008484 | 0.3649 | 0.7522 | 1.418 | 0.4937 | 0.1779 | 0.005276 | 0.227 | 0.4678 | 0.8816 |
| 2011 | 0.3684 | 0.1328 | 0.003938 | 0.1694 | 0.3491 | 0.6579 | 0.8512 | 0.3068 | 0.009097 | 0.3913 | 0.8065 | 1.52 | 0.5293 | 0.1908 | 0.005657 | 0.2433 | 0.5016 | 0.9452 |
| 2012 | 0.3748 | 0.1351 | 0.004006 | 0.1723 | 0.3552 | 0.6693 | 0.8659 | 0.3121 | 0.009255 | 0.3981 | 0.8205 | 1.546 | 0.5385 | 0.1941 | 0.005755 | 0.2476 | 0.5103 | 0.9616 |
| 2013 | 0.37 | 0.1334 | 0.003954 | 0.1701 | 0.3506 | 0.6607 | 0.8547 | 0.3081 | 0.009135 | 0.3929 | 0.8099 | 1.526 | 0.5316 | 0.1916 | 0.005681 | 0.2444 | 0.5037 | 0.9492 |
| 2014 | 0.3588 | 0.1293 | 0.003835 | 0.165 | 0.34 | 0.6408 | 0.829 | 0.2988 | 0.00886 | 0.3811 | 0.7855 | 1.48 | 0.5155 | 0.1858 | 0.00551 | 0.237 | 0.4885 | 0.9206 |
| 2015 | 0.349 | 0.1258 | 0.00373 | 0.1604 | 0.3307 | 0.6232 | 0.8062 | 0.2906 | 0.008617 | 0.3706 | 0.7639 | 1.44 | 0.5014 | 0.1807 | 0.005359 | 0.2305 | 0.4751 | 0.8953 |
| 2016 | 0.36 | 0.1298 | 0.003848 | 0.1655 | 0.3411 | 0.6429 | 0.8317 | 0.2998 | 0.008889 | 0.3824 | 0.7881 | 1.485 | 0.5173 | 0.1864 | 0.005528 | 0.2378 | 0.4901 | 0.9237 |

Annex 3. Silky shark estimates in number (million sharks) for the target species catch-based estimation method


Annex 4. Silky shark estimates in number (million sharks) for the longline effort-based estimation method


Annex 5. Silky shark estimates in biomass (in ' 000 t ) for the area-based estimation method


Annex 6. Silky shark estimates in biomass (in ' 000 t ) for the target species catch-based estimation method




[^0]:    ${ }^{1}$ Technical Coordinator-Sharks and Bycatch, Common Oceans (Areas Beyond National Jurisdiction (ABNJ)) Tuna Project, Food and Agriculture Organization of the United Nations, Rome, Italy

[^1]:    ${ }^{1}$ The FAO chondrichthyan capture production figures include rays and chimaeras as well as sharks, but it is appropriate to include rays and chimaeras because their fins are also used in the shark fin trade.

