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Synthesizing a network of evidence on a seabird bycatch mitigation measure

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Synthesizing a network of evidence on a seabird bycatch mitigation measure

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20 ABSTRACT

USA

- 21 Robust estimates of the relative efficacies of alternative management interventions are essential
- 22 for developing evidence-informed fisheries bycatch policy. Bycatch is a major threat to the
- conservation of albatrosses and other pelagic seabirds. Branchline weighting is one approach
- 24 prescribed by regional fisheries management organizations and the Agreement on the
- 25 Conservation of Albatrosses and Petrels to reduce seabird bycatch in pelagic longline fisheries.
- 26 We used a Bayesian multilevel network meta-regression modelling approach to conduct the first
- 27 synthesis of available evidence to assess the relative efficacies at mitigating seabird bycatch of
- 28 alternative pelagic longline weighting designs. Unlike conventional pairwise meta-analysis,
- network meta-analysis enables the simultaneous comparison of multiple interventions within a
 coherent modelling framework. There was a >97% probability that all weighting designs
- 31 significantly reduced seabird bycatch compared to a reference design with no weight within 5m
- 32 of the hook. Nonetheless, some weighting designs were significantly more effective at reducing
- 33 seabird bycatch than others for instance, the 2 designs with weights >60g and >1m from the
- hook performed the best with >93% probability that those 2 designs performed significantly
- 35 better than 2 more commonly used designs with less weight but attached closer to the hook.
- 36 These two best performing designs reduced seabird bycatch by *ca.* 89% relative to the
- 37 reference design. These relative efficacies and rankings, when combined with other
- performance criteria such as costs to commercial viability and crew safety, support robust
 evaluations of alternative bycatch management strategies.
- 40
- Keywords: Bayesian network meta-analysis; Bycatch mitigation; Evidence synthesis; Fishing
 gear weights; RFMO; Seabirds
- 43 44

45 **1. INTRODUCTION**

46 Incidental capture in fisheries is a major threat to the conservation of pelagic seabirds and

47 reduces fishing efficiency (Brothers et al., 1999; Anderson et al. 2011; Phillips et al., 2016). This

- 48 mortality source is particularly problematic for albatrosses and petrels, which are two of the
- 49 three most threatened groups of seabirds (Phillips et al. 2016; Dias et al. 2019). Of the suite of
- 50 anthropogenic threats that albatrosses and petrels are exposed to, fishing-associated mortality
- 51 causes the largest population impacts (Heppell et al., 2000; Bakker et al., 2018). For depleted

52 populations, even low fishing-associated mortality levels can have profound effects and risk 53 extirpation (Heppell et al. 2000).

54 Pelagic longlining is a main fishing gear type supplying canned, frozen and fresh 55 products from tuna and tuna like species (Scombroidei), billfishes (Xiphioidei) and other large 56 pelagic and neritic apex- and meso-predators (Miyake et al., 2010; Williams and Ruaia, 2022). 57 Primarily at higher latitudes, pelagic longline fisheries are a major hazard to exposed pelagic 58 seabirds (Anderson et al. 2011; Dias et al. 2019). Several methods can substantially reduce 59 seabird bycatch rates in pelagic longline fisheries that cause minimal or no costs to economic viability, practicality and crew safety (Hall et al., 2017; ACAP 2023). However, there has been 60 mixed progress in the adoption of seabird bycatch mitigation measures by fisheries 61 62 management authorities and compliance is uncertain due to deficits in fisheries monitoring, 63 surveillance and enforcement (Juan-Jorda et al., 2018).

64 Branchline weighting is one option for seabird bycatch mitigation in pelagic longline 65 fisheries, which is typically prescribed to be used in combination with other seabird bycatch mitigation methods. Some pelagic longline vessels conventionally incorporate lead-centered 66 67 weighted swivels into branchlines of various mass and locations to: increase the gear sink rate; 68 maintain the gear at a desired fishing depth; affect the action (movement) of the terminal baited 69 hooks; and reduce line tangles (Brothers et al., 1999; Beverly et al., 2003). The mass of pelagic 70 longline branchline weights and distance between the weight and hook significantly affect 71 seabird catch risk during setting (Robertson et al., 2013; Melvin et al., 2013, 2014; Santos et al., 72 2019), as well as during the gear soak and haul (Gilman et al., 2014b; Phillips and Wood, 2020). 73 During setting, these two variables affect the sink rate of baited hooks and their availability to 74 seabirds. The initial baited hook sink rate affects seabird catch risk near the surface. The sink 75 rate of a baited hook begins to be affected by a branchline weight once the weight sinks to a depth when the leader (the section of the branchline between the hook and weight) becomes 76 77 taut (Robertson et al., 2010). The closer a weight is to the hook and the larger the mass, the 78 less time it takes for the weight to affect the hook's sink rate near the surface and the less time 79 baited hooks are available to seabirds to the extent of their diving depth (Robertson et al., 2013; Jimenez et al., 2019). Similarly, during the gear soak, in fisheries where hooks fish near the 80 surface, in deeper-set fisheries when catch bring the gear to the surface, as well as when the 81 82 gear shoals from currents and wind, and during the haul when hooks are brought back to the 83 surface, the closer a weight is to the hook, the more likely the weight will reduce the availability 84 of baited hooks to surface-foraging seabirds (Gilman et al., 2014b).

85 In regions where secondary seabird interactions occur or where primary interactions 86 occasionally occur at depth, the hook sink rate to depths attainable by these deep-diving 87 species is an additional important predictor of seabird catch risk (Jimenez et al., 2012; Melvin et al., 2014; Rollinson et al., 2016; Guilford et al., 2022). Secondary interactions occur when deep-88 89 diving seabirds access baited hooks at depth and return them to the sea surface, providing 90 larger primarily surface-foraging seabird species with a second opportunity to access baited 91 hooks (Jimenez et al., 2012; Melvin et al., 2014). For fisheries in these regions, the effect of the 92 branchline weighting design on the baited hook sink rate both near the surface where most or all primary interactions occur and to as deep as ca. 25m for secondary interactions by proficient 93 94 divers such as species of shearwaters and *Procellaria* petrels (Bentley et al., 2021; Rollinson et al., 2016; Frankish et al., 2021) will affect seabird catch risk. In some regions, such as the 95 central north Pacific Ocean, secondary interactions are rare events and seabirds susceptible to 96 97 capture typically only make body thrusts to reach prey within ca. 1 m of the surface (Prince et al., 1994; Kazama et al., 2019; Bentley et al., 2021, Guilford et al., 2022). For fisheries in these 98 99 regions, the effect of the branchline weighting design on the sink rate very near the surface has 100 a greater effect on seabird bycatch risk than the sink rate to depth.

101 Unlike many other seabird bycatch mitigation methods, robust compliance monitoring of 102 a required branchline weighting design when conventional lead-centered swivels are used, which are crimped in place, is feasible by dockside inspections for vessels making relatively
short trips, where it would be impractical for crew to move weights during a fishing trip (ACAP,
2023). This is an important consideration given the extremely limited onboard monitoring that

106 occurs in global pelagic longline fisheries (Gilman et al., 2014a). Surveillance of line weighting

designs of vessels making long trips would, however, require onboard observers (and possibly
 by electronic monitoring systems, Gilman et al., 2019). The location of attachment of sliding

109 weights and weighted hook shielding devices can be easily changed by crew during both short

and long trips, making dockside inspections an ineffective compliance monitoring approach.

111 This could potentially be addressed by requiring unweighted swivels to be attached at the

maximum permitted distance that weights can be attached from hooks.

113 Five regional fisheries management organizations (RFMOs) have adopted binding 114 measures that include branchline weighting as one option to meet seabird bycatch mitigation 115 requirements by pelagic longline fisheries in designated areas, where each RFMO defines three 116 or four options for weighting designs (IATTC, 2011; ICCAT, 2011; WCPFC, 2018; SIOFA, 2022; 117 IOTC. 2023). The Agreement on the Conservation of Albatrosses and Petrels (ACAP) 118 recommends three weighting design options (ACAP, 2023). ACAP's (2023) recommendations 119 are based on mechanistic studies that measured sink rates of alterative branchline weighting 120 designs when considering whether the estimated sink rates are predicted to afford adequate 121 protection when a vessel is using a bird-scaring tori line and considering seabird diving depths. 122 The ACAP (2023) recommendations are also based on findings from three selected studies on 123 the effects of alternative weighting designs on seabird interaction rates (Gianuca et al. 2011; 124 Barrington et al., 2016; Jiménez et al. 2019a; Santos et al. 2019).

125 The controlled experiments have a higher strength of evidence compared to sink rate 126 mechanistic studies, while findings from meta-analytic synthesis studies have the highest strength of evidence, producing the most robust and generalizable findings that are optimal for 127 128 informing global and regional seabird bycatch policy (Gilman and Chaloupka, 2023). A synthesis 129 of accumulated evidence of seabird bycatch rate responses to alternative branchline weighting 130 designs, including those prescribed by RFMOs and ACAP, has not previously been conducted. 131 Meta-analytic syntheses, which synthesize estimates sourced from a mixture of independent and context-specific studies, have the lowest risk of error and bias and the highest strength of 132 133 evidence, producing the most robust and generalizable findings that are optimal for evaluating 134 alternative regional bycatch management strategies (Dicks et al. 2014; Nichols et al. 2019).

135 This study provides the strongest evidence to guide regional policies on branchline 136 weighting to mitigate seabird bycatch. The study objective was to determine whether certain 137 pelagic longline branchline weighting designs, including those required by RFMOs and 138 recommended by ACAP, have similar or otherwise significantly different seabird catch rates. We 139 also explored informative predictors of seabird bycatch risk other than branchline weighting. We 140 conducted a systematic literature review to compile relevant publications and assembled a 141 dataset suitable for meta-analytic evaluation. We used a Bayesian multilevel network meta-142 regression modelling approach to synthesize available evidence to assess the relative efficacies 143 of branchline weighting designs that are used in pelagic longline fisheries to mitigate seabird 144 bycatch. A key outcome of the network meta-synthesis is a league table or rankings of those 145 relative efficacies that enable identifying the least to most efficacious designs. Findings can be 146 used by fisheries management authorities to define evidence-informed weighting designs that 147 meet objectives for managing seabird bycatch in pelagic longline fisheries.

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150 **2. METHODS**

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152 **2.1. Data Compilation**

153 A two-tiered systematic and unstructured literature search was employed to compile relevant

- 154 peer-reviewed published and grey literature, including government observer program data, and
- 155 to assemble a dataset suitable for meta-analytic evaluation. The methods for the systematic 156 literature review were adapted from the Reporting Standards for Systematic Evidence
- 157 Syntheses (Haddaway et al., 2018), Collaboration for Environmental Evidence (Pullin et al.,
- 158 2020), and extended PRISMA framework applicable for network meta-analyses (NMA) (Hutton
- et al 2015). Details on the literature review methods, including searched databases, search
- 160 strings, study eligibility criteria and screening process for retention, and extracted data fields are
- 161 in Supplemental Material Section S1. There were 12 retained publications obtained from the
- systematic review (Supplemental Material Fig. S1, Table S2). One additional retained
- publication (Table S2) and five retained observer program datasets from New Zealand, Taiwan,
 the U.S. Atlantic Ocean, the Hawaii-based U.S. Pacific Ocean shallow-set swordfish fishery.
- the U.S. Atlantic Ocean, the Hawaii-based U.S. Pacific Ocean shallow-set swordfish fishery,
 and the Hawaii-based U.S. Pacific Ocean deep-set tuna fishery (Supplemental Material Section)
- 166 S4) were obtained from the unstructured literature review.
- 167

168 **2.2. Branchline Weighting Design Treatments**

169 Supplemental Material Table S1a summarizes the branchline weighting designs prescribed by 170 five RFMOs with binding seabird measures and recommended by ACAP. To produce a 171 manageable number of weighting design treatments and with adequate sample sizes we used 172 only the \geq designs (e.g., the prescribed design \geq 45g weights within 1m of the hook was retained 173 and eliminated >45g within 1m). As none of the compiled records had >1 weight within half a 174 meter of the hook, we combined the design prescribed by one of the RFMOs that specifies this 175 limit with an ACAP recommended design of ≥40g within 0.5m. We adapted the RFMO/ACAP 176 designs to provide more detailed definitions so that a single branchline weighting design can fall under only one unique category. An RFMO-prescribed designs of ≥98g attached between 177 178 >3.5m and ≤4m of the hook was excluded because there was only a single record in this 179 category. We included two categories not prescribed by RFMOs or ACAP. This resulted in the 180 following 7-category treatment variable for branchline weighting design for the modeling, where 181 A is referred to as the reference design and is not prescribed by RFMOs or ACAP, designs B 182 through F are prescribed RFMO or ACAP options (sample sizes are number of records), and 183 design G is not prescribed by RFMOs or ACAP:

- 184
- 185 A. 0g attached \leq 5m from the hook (N=86)
- 186 B. \geq 40g attached \leq 0.5m from the hook (N=42)
- 187 C. \geq 45g to <60g between >0.5m and \leq 1m from the hook (N=37)
- 188 D. $\geq 60g$ between > 0.5m and $\leq 1m$ from the hook (N=27)
- 189 E. $\geq 60^\circ$ to <80g between >1m and $\leq 3.5m$ or $\geq 80^\circ$ g between >2m and $\leq 3.5m$ from the hook 190 (N=18)
- 191 F. \geq 80g between >1m and \leq 2m from the hook (N=9)
- 192 G. \geq 8g to \leq 39g between \leq 1m or \geq 45g to \leq 80g between \geq 1.3m and \leq 4m from the hook 193 (N=52)
- 194
- Table S1b identifies which of the combined RFMO and ACAP design options are met by each of the above 7 branchline weighting design treatments included in the NMA. Table S3 summarizes the records covered by each of the 7 treatments, including the number of fisheries and regions from which they were derived, and the range and median weight mass and leader lengths of the records within each treatment.
- 201 **2.3. Statistical Modelling Approach**

202203 2.3.1. Visualising the network geometry

204 We summarise the underlying NMA structure for our study in Figure 1, which comprises the 7 205 pelagic longline branchline weighting designs (A-G) with 21 weighting design combinations or interactions — this plot is a mandatory component of any network modelling workflow (Liu et al 206 207 2023). Recently, Freeman et al (2023) also proposed the use of upset plots (Conway et al 2017) 208 for further exploring and visualising the NMA dataset characteristics, which we implemented 209 using the ComplexUpset R package (Krassowski 2020) for the weighting design combinations 210 shown in the network geometry plot (Figure 1). The upset plot display of the network structure is 211 shown in Figure S2.

212

213 2.3.2. Statistical modelling framework

NMA is a well-established statistical modelling approach used in the healthcare evaluation, pharmacometrics, psychology and clinical epidemiology domains for comparing the relative efficacies of three or more interventions or treatments by synthesising aggregate and/or individual participant evidence sourced from multiple trials (Ades et al 2024). An accessible introduction to Bayesian NMA can be found in Liu et al (2023), the comparative benefits of Bayesian inferencebased meta-analysis can be found in (Pappalardo et al 2020), while an authoritative primer on Bayesian statistical modelling in general can be found in van de Schoot et al (2021).

221 Conventional pairwise meta-analysis is based on synthesising direct evidence 222 comparing only two treatments from multiple trials that can only generate so-called unrelated 223 treatment effects. For instance, say there are three possible treatments A-C but only AB and AC comparisons are known from direct head-to-head clinical trials - no BC trials have ever 224 225 occurred and so any inference about such an effect remains unknown in conventional pairwise 226 meta-analysis. NMA leverages both direct and indirect evidence for three or more treatment 227 comparisons, so the BC effect can be inferred indirectly as part of a connected network of trials 228 based on direct evidence from the AB and AC trials - a property known as coherence (Ades et 229 al 2024). The relative rankings of those NMA-based treatment effects can then be used to 230 identify the "best" treatment, given that specific set of treatments, and to support evidence-231 informed cost-effectiveness treatment recommendations given those rankings.

232 Specifically, we used a Bayesian multilevel (or random-effects) network meta-regression 233 modelling approach (Phillippo et al 2020) to synthesis all available studies comparing longline 234 aear weighting designs on seabird bycatch rates (Figure 1). Importantly, not all seven weighting 235 designs need to be in each study so long as each study comprises a subset of at least two or 236 more of the designs. We could only source the study-specific summarised or aggregate data 237 and not the pelagic longline set-specific data for each study, although a mix of both data 238 sources, if available, can be combined in a single integrated Bayesian multilevel network meta-239 regression model (Phillippo et al 2020, Singh et al 2022).

240 We also sourced a range of study-specific marginal covariate summaries that we could 241 then use as potentially informative seabird bycatch predictors in our network meta-regression 242 modelling framework. See Supplemental Material Sections S1 and S2 for a description of 243 potentially informative predictors that were extracted from retained publications and explored for 244 potential inclusion in the models. The explored predictors included latitude, bait type, hooks 245 between floats, oceanic region and five binary predictors for the following additional seabird 246 bycatch mitigation measures: tori line, night setting, side setting, hook shielding device, and 247 blue-dyed bait (Supplemental Material Table S2).

We fitted our multilevel network meta-regression models using the Stan computation back-end (Carpenter et al 2017) via the multinma interface for R (Phillippo 2024) with Poisson likelihood (log link) appropriate for study-specific aggregate seabird catch data and with the study-specific aggregate number of hooks (fishing effort) as the exposure metric. The randomeffects structure (intercepts only) comprised the study-specific identities. These models are therefore Poisson likelihood GLMMs with the response or effect size being the study-specific log rate ratio (Dias et al 2011, Hooper et al 2000) – also known as the log response ratio (Lajeunesse 2011). We also used QR-decomposition and numeric regression term centering in
 our network meta-regression models, fitted using multinma (Phillippo 2024), to improve
 computational efficiency by guarding against potential multicollinearity among covariates that
 might affect the model posterior geometry (Betancourt 2017).

259 All models were fit with 4 Markov chains with 5,000 iterations per chain after a warm-up of 260 1000 iterations. Model convergence was assessed using parameter-specific diagnostics such as 261 multiple chain rank plots, bulk and tail effective sample size metrics and a rank-based Rhat 262 statistic (Vehtari et al 2021). All inference was based on the 20,000 network meta-regression 263 model posterior draws, which were accessed as multidimensional arrays to support a range of 264 post-processing functions provided by the bayesplot (Gabry et al 2019) and tidybayes R 265 packages (Kay 2023). For instance, those posterior draws (samples) were used to derive the 266 uncertainty intervals (HDIs or highest posterior density intervals: Kruschke and Liddell 2018; and 267 see also Pappalardo et al 2020). The ggplot2 (Wickham 2016), ggdist (Kay 2024) and 268 colorspace (Zeileis et al 2020) packages for R were used for all the summary graphics while 269 the patchwork R package (Pedersen 2020) was used for all multi-panel arrangements.

270

271 **2.3.3. Relative ranking of weighting designs**

272 Using a Bayesian inference framework for our network meta-synthesis of the 7 branchline 273 weighting designs (or treatments) enables us to derive the full posterior probability distribution of 274 all the relative weighting design effects and hence robust summaries of uncertainty of those 275 effects. Additionally, this framework supports the robust ranking of those relative treatment 276 effects in the network, which is especially helpful for supporting evidence-informed bycatch 277 mitigation policy. In a Bayesian modelling framework, the ranks are derived from the effect-278 specific posterior distributions using two approaches based on: (1) the mean or median of the 279 posterior distribution of the ranks with the 95% credible interval (Phillippo et al 2020), or (2) the 280 probability of a specific treatment ranking the best determined as the Surface Under the Cumulative RAnking curve (SUCRA: Veroniki et al 2018, Daly et al 2019). The SUCRA 281 probabilities are appealing in terms of the ease of its interpretation and Liu et al (2023) propose 282 283 that rank probabilities > 0.90 reflect an efficacious treatment effect. On the other hand, the 284 posterior ranking approach is more robust than SUCRA to underlying model assumptions. We 285 used both approaches here (posterior median ranking, SUCRA).

286

287 **2.3.4. Model selection using cross-validation**

Model selection procedures were based on leave-one-out cross-validation (LOOcv) metrics to estimate any comparative difference in expected predictive accuracy between the various models fitted (Vehtari et al 2017) — such as whether the model included covariates or not. The weight of evidence in favour of one model over any other candidate models was also assessed using Bayesian stacking, which is the Bayesian analogue of model averaging (Yao et al 2018).

293

294 **2.3.5.** Assessing network model assumptions

According to Seide et al (2020), there are three fundamental assumptions in NMA: (1) transitivity, (2) network consistency and (3) network heterogeneity (see also Ades et al 2024 and the concept of exchangeability). Not all our studies comprised all the 7 weighting designs in head-to-head comparisons (see upset plot, Figure S2), so we assume that any missing designs within a study are missing at random — there are no consensus statistical tests of this specific assumption, but it is conceptually related to the following consistency assumption (Seide et al 2020).

The all-important consistency assumption concerns whether the direct (head-to-head) estimates and the indirect estimates (made by inference) within each contrast in the connected network are sufficiently similar. This assumption can be readily evaluated using either so-called

305 global assessment tests (unrelated mean effects or UME model: Spineli 2022) and local 306 assessment tests (node-splitting procedure: Veroniki et al 2013) — we used both forms of consistency assessment tests applied to our best-fit Bayesian network meta-regression model 307 308 used for inference. The global assessment was done by comparing the expected predictive 309 accuracy between the best-fit model and the unrelated mean effects model using leave-one-out 310 cross-validation (see `Model selection` section above). The UME model is based only on direct 311 evidence and estimates each treatment comparison as a separate and unrelated parameter 312 (Spineli 2022).

The local consistency assumption was specifically assessed using visual review of nodesplitting plots (Currier et al 2023) derived using the multinma package for R (Phillippo 2024). The heterogeneity assumption is especially important for small-sample studies (Ades et al 2024) and can be readily explore using prior sensitivity analysis, which we implement in our modelling workflow by fitting meta-regression models with weakly informative priors (Lemoine et al 2019) and then using more informative priors recommended for various forms of the between-study heterogeneity parameter (Turner et al 2015, Seide et al 2020).

320

321 **2.3.6.** Weighting design-specific relative risk reduction in seabird bycatch

We back-transformed the predicted seabird bycatch log relative risk contrasts derived for the 6 322 323 branchline weighting design treatments (B-G) relative to reference design A (no weight ≤5m 324 from the hook) using the transform-then-summarize posterior samples approach since such 325 summary metrics are not transform invariant (Wang et al 2018). We then used these back-326 transformed posterior samples to calculate the mean percent relative reduction in seabird 327 bycatch risk (Gilman et al 2020) for the 6 branchline weighting design treatments relative to the 328 reference design. As in Gilman et al (2020), we also used the highest posterior density interval 329 (HDI) as our measure of uncertainty (Kruschke and Liddell 2018). The mean and 95% HDIs 330 were summarized from the posterior samples for each design relative to the reference design 331 using the tidybayes package for R (Kay 2023).

332 333

334 **3. RESULTS**

335

336 **3.1. Model Evaluation Diagnostics**

337 Leave-one-out cross-validation (LOOcv) and Bavesian stacking metrics suggested that the best-338 fit model fitted to the pelagic longline weighting design network geometry (Figure 1) was a meta-339 regression model that included predictors for oceanic region and the 5 seabird bycatch 340 mitigation measures (looic difference = -105.9, se = 75.9; ca. 98% weight of evidence in favour 341 of model with covariates) — there was little support for a NMA model without the covariates. 342 Convergence diagnostics such as multiple chain rank plots, and the effective posterior sample 343 size (ESS) metrics coupled with the rank-based diagnostic statistic Rhat < 1.01 (Vehtari et al 344 2021), all reflected convergence of the Bayesian network meta-regression model with Poisson likelihood that was used here for statistical inference. Nonetheless, residual deviance for all 345 346 models fitted here was greater than the number of data points, which can suggest that there 347 might be grounds for further improvement in model fit (Phillippo et al 2020) — see `Covariate 348 effects` section below.

349

350 3.2. Prior Sensitivity

351 We assessed comparative predictive accuracy of three between-study heterogeneity parameter

- priors for the meta-regression model: (1) a weakly informative *half-normal*(scale=5²) prior on the
- heterogeneity parameter standard deviation or τ (Phillippo 2024), (2) a more informative *half-*
- *normal*(scale=0.5) prior on the heterogeneity variance (τ^2) proposed by Seide et al (2020) on

355 theoretic grounds, and (3) a more informative empirically-derived lognormal(-3.93, 1.51²) prior 356 on τ^2 suggested by Turner et al (2015). LOOcv and Bayesian stacking metrics suggested there 357 was no meaningful difference in predictive accuracy between a network meta-regression model 358 with the weakly informative heterogeneity prior and the same meta-regression model but now 359 with a more informative *lognormal* heterogeneity prior (looic difference = -1.6, se = 1.9) — there 360 was no support for the informative prior suggested by Seide et al (2020). There is a strong 361 empirical basis for the *lognormal* heterogeneity prior (Turner et al 2015), so we proceed with 362 inference based on our network meta-regression model with that prior on the heterogeneity 363 parameter variance.

364

365 3.3. Weighting Design Specific Effects

366 The predicted design-specific conditional effects relative to the reference design (A) based on 367 the best-fit Bayesian network meta-regression model fitted to the weighting designs network 368 structure (Figure 1) are summarised in Figure 2. This summary display shows both the full 369 posterior distribution of each parameter coupled with key summary metrics (median, highest 370 posterior density intervals) that helps to support a more informative means of communicating 371 parameter uncertainty (van der Bles et al 2019). The six branchline weighting designs B-G are 372 all significantly more effective at reducing expected seabird bycatch than weighting design A -373 the predicted effects are all significantly < 0 on the log response rate scale. Even for design G 374 we can be > 97% sure that the predicted effect is < 0 (in other words, > 97% of the posterior density shown in Figure 2 is < 0). We can be near 100% sure of this reduced seabird bycatch 375 376 effect for design E and *ca.* 99% sure for design F. The largest relative effect is for design F but 377 there is greater uncertainty associated with that effect due to the smaller sample size for design-378 specific comparisons that included F (see Figure 1, Figure S2). Given this uncertainty, it is 379 apparent that weighting designs E and F appear to have similar efficacy for reducing seabird 380 bycatch in pelagic longline fisheries.

381

382 3.4. Covariate Effects

Figure S3 summarises the conditional effects for the covariates or predictors of seabird bycatch rate included in our best-fit network meta-regression model. The most prominent conditional effects were for (1) the apparent decreased seabird bycatch risk for pelagic longline fisheries operating in the South Pacific region compared to fisheries in the North Atlantic region, and (2) the apparent decreased seabird bycatch risk given the use of night-setting or side-setting as seabird mitigation measures. Note that all covariates were study-specific marginal covariate summaries of aggregate data and not based on set-specific predictors.

390 Despite inclusion of these covariates as main-effects, there remained moderate 391 between-study heterogeneity, by comparison with expected treatment effects (Figure 2, see 392 also Phillippo et al 2020), that was unaccounted for by our best-fit network meta-regression 393 model (posterior mean τ = 2.02; 95% quantile-based CI: 1.67-2.5). High between-study 394 heterogeneity is an expected and common finding for nonexperimental study-based meta-395 analyses (Turner et al 2015). The between-study heterogeneity apparent in our study might be 396 considerably reduced if we had included covariate interaction terms with say the weighting 397 designs in our models but such interactions were precluded due to unbalanced covariate or 398 effect modifier combinations (Phillippo et al. 2020) — for instance, not all 7 designs were 399 deployed across all 5 ocean regions and so precluded an ocean*design interaction term.

400

401 3.5. Checking for (In)Consistency

402 The global consistency test compared the predictive accuracy of an unrelated mean effects or

- 403 UME model (Spinelli 2022) with our best-fit seabird bycatch meta-regression or consistency
- 404 reference model (see Figure 2). The LOOcv and Bayesian stacking metrics suggested that both

406 evidence in favour of the best-fit model compared to the UME model). This implies that there

407 was no evidence of global inconsistency and hence our network meta-regression model was fit

408 for inference. Local consistency was evaluated here using the network node-splitting approach 409 and compared the direct and indirect evidence of all pairwise contrasts (node-splits) with

and compared the direct and indirect evidence of all pairwise contrasts (node-splits) with
 sufficient data (15 of the 21 possible contrasts). The direct and indirect posterior density

410 sumplem data (15 or the 21 possible contrasts). The direct and indirect posterior density 411 distributions for those 15 node-splits are summarised in Figure S4, which shows no evidence of

- 412 inconsistency. There was no evidence for network model inconsistency based on either the
- 413 global or local tests.
- 414

415 3.6. Efficacy Ranking

416 The median posterior rank for each weighting design based on the posterior samples sourced 417 from our best-fit meta-regression model is summarised in Figure 3. Importantly, there was > 418 97% probability that all weighting designs significantly reduced seabird bycatch compared to 419 reference design A with no weight attached within 5m of the hook. Nonetheless, some weighting 420 designs were significantly more effective at reducing seabird bycatch risk. Designs E and F rank 421 the "best", although there is considerable uncertainty associated with design F due to small 422 sample size, which is a concern for robust treatment rank scoring (Daly et al 208, Ades et al 423 2024). An alternative ranking approach uses the model-based cumulative rank order 424 probabilities based on the SUCRA curve shown in Figure S5. The posterior ranking plot (Figure 425 3) and the SUCRA-based ranking plot (Figure S5) show very similar expected weighting design 426

rankings — the only difference being a re-order between designs C and B that are in close
 proximity in rankings and supports the conclusion that the overall rank ordering is robust (Daly

- 428 et al 2018). Overall, the two best performing designs (E and F) were found to reduce expected
- 429 seabird bycatch by *ca.* 89% relative to the reference design: E v A ~ 88.8% (95% HDI: 72-99)
- 430 and F v A ~ 88.9% (95% HDI: 63-99). The next best performing design D reduced expected

431 seabird bycatch by *ca.* 76% relative to the reference design: D v A ~ 76.1% (95% HDI: 43-97).

432

433 3.7. Relative Effect Contrasts

434 The relative design-specific effects for all 21 pairwise contrasts between all 7 branchline 435 weighting designs based on the best-fit Bayesian network meta-regression model are 436 summarised in Figure 4 — the first six contrasts repeated the main effects summarised in Figure 2. There was a > 98% probability that all 5 RFMO and ACAP prescribed weighting 437 designs (B-F) had lower expected seabird bycatch rates relative to the reference design A. Even 438 439 design G (not a prescribed RFMOs or ACAP design) was estimated to reduce the expected 440 seabird bycatch rate by ca 60% (95% HDI: 7-96) compared to the reference design. The 441 contrast estimates in Figure 4 are especially useful for considering design effects that might be 442 more cost-effective or more practical in terms of crew safety when deploying pelagic longline 443 fishing gear. Using the E vs B contrast as an example, there is a > 97% probability that design E 444 is more effective at reducing seabird bycatch than design B. Or there is a > 95% probability that 445 design E is more effective at reducing seabird bycatch than either designs B or C. On the other 446 hand, there is a > 97% probability that design G is more likely to increase seabird bycatch 447 compared to design E or a > 94% probability that G is more likely to increase seabird bycatch 448 compared to design F.

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451 **4. DISCUSSION**

452453 **4.1. Relative Efficacies of Alternative Weighting Designs**

454 Unlike conventional pairwise meta-analysis, NMA enables the simultaneous comparison of

- 455 multiple treatments or interventions within a coherent modelling framework (Ades et al 2024).
- 456 NMA approaches achieve this by leveraging direct within-study evidence and indirect across-

studies evidence for all treatments linked in a connected network structure — even if some
treatments have not been tested in direct or head-to-head comparisons. The foremost product
of any NMA is a league table with rank-ordering of all the treatments to help identify for a
specific outcome the relative efficacies of treatments within that network.

461 Our Bayesian multilevel NMA modelling approach synthesized all available evidence to 462 assess the relative efficacies of seven branchline weighting designs for mitigating seabird 463 bycatch in pelagic longline fisheries. The resulting league table identifies the most effective 464 designs. Six treatments had significantly lower expected seabird bycatch rates than the 465 reference design with no weight within 5m of the hook, consistent with the expectation that 466 attaching weights far from the hook results in relatively slow baited hook sink rates, which 467 increases seabird bycatch risk (Brothers et al., 1999; ACAP, 2023). The five RFMO and ACAP 468 prescribed designs had lower expected seabird bycatch rates relative to the reference design, 469 ranging from *ca.* 66% (design B, 95% HDI:21-95) to *ca.* 89% (design F, 95% HDI: 63-99) lower 470 expected seabird bycatch rates. There was a > 98% probability that all 5 RFMO and ACAP 471 prescribed weighting designs had lower expected seabird bycatch rates relative to the reference 472 design. The one assessed design not prescribed by RFMOs or ACAP was estimated to reduce 473 the expected seabird bycatch rate by ca. 60% (95% HDI: 7-96) compared to the reference 474 design, and we can be > 97% sure that design G has a lower seabird bycatch rate than the 475 reference design.

476 RFMOs and ACAP prescribe multiple options for branchline weighting designs based on 477 the hypothesis that the alternative designs produce similar seabird catch rates. This is an 478 implicit hypothesis of RFMO seabird measures (IATTC, 2011; ICCAT, 2011; WCPFC, 2018; 479 SIOFA, 2022; IOTC, 2023) and explicit in the ACAP (2023) recommendations. Some of the line 480 weighting designs included in RFMO seabird measures may have been derived from member 481 delegations who ensured that contemporary practices of their fisheries were included as options 482 (e.g., WCPFC, 2006). The ACAP (2023) recommended line weighting design options were 483 defined based on mechanistic studies that identified weighting designs that produce hook sink rates that were predicted to adequately protect seabirds when used in combination with a tori 484 485 line, and based on findings from three experiments on branchline weighting design effects on 486 seabird interaction rates (ACAP, 2023). Our study findings do not support this hypothesis, as we 487 found that some designs prescribed by RFMOs and ACAP have significantly different seabird 488 catch risks. Of 10 possible pairwise contrasts of the 5 prescribed weighting designs, 4 pairs (E 489 vs. B, E vs. C, F vs. B, and F vs. C) had >92% probabilities of having significantly different 490 seabird bycatch rates. The non-prescribed weighting design included in the NMA models had 491 similar efficacy as two of the prescribed designs. Following the reference treatment, these 3 492 designs had lowest posterior ranks, and had significantly higher seabird catch rates relative to 493 the two designs with lowest seabird catch rates.

494 One weighting design prescribed by all 5 RFMOs with binding seabird management 495 measures (Table S1) of ≥98g attached between >3.5m and ≤4m of the hook had too small a 496 sample size for inclusion in the meta-analysis and has an unknown relative efficacy. Discussed 497 in more detail below, the study was limited due to small sample sizes for some prescribed 498 weighting designs by region and number of fisheries. With more certain estimates of the 499 efficacies of alternative prescribed weighting designs, those with relatively low predicted 500 efficacies could be eliminated to augment seabird conservation. Or, for two RFMOs with seabird measures with two lists of methods hypothesized to be relatively more and less effective 501 502 (IATTC, 2011; WCPFC, 2018), weighting designs could be split into these two lists based on 503 their predicted relative efficacies.

504 Figure 3 (see also Figure S4) shows the estimated relative efficacies of the 7 designs 505 based on an outcome criterion measured as seabird bycatch. Other key performance criteria to 506 holistically evaluate alternative strategies' simulated outcomes across the multiple populations 507 and stocks of exposed commercial and at-risk bycatch species are described in Supplemental Material Section S8. The design-specific contrasts shown in Figure 4 can be evaluated in
association with these other management-focused performance criteria. Moreover, the
confidence in any guidelines or recommendations based on such branchline weighting designspecific contrasts can then be assessed quantitatively within a Bayesian NMA framework using
techniques such as threshold analysis (Phillippo et al 2019). Threshold analysis quantifies how
much the evidence would need to change before a management recommendation would be

514 revised, and an assessment could then be made to predict whether the evidence could plausibly 515 change to a degree that exceeds this threshold (Phillippo et al 2019).

516

517 4.2. Effects of Other Seabird Bycatch Mitigation Methods and Ocean Region

518 The apparent differences in relative seabird bycatch rates of some of the five bycatch mitigation 519 methods included as model covariates might be partly explained by the effect of crew behavior 520 on the efficacy of some methods. Most (90%) of records in the assembled dataset fit to the NMA 521 models were derived from observer program datasets. Efficacies of tori lines, blue-dyed bait and 522 side setting can be strongly affected by crew behavior. As a result, they might be less effective 523 in practice, during real-world, commercial fishing operations, as determined through analyses of 524 observer and electronic monitoring data, than findings from experiments, despite the latter 525 having a relatively lower risk of error and bias (Gilman and Chaloupka, 2023). Unlike in 526 experimental studies, observational studies do not experimentally manipulate specific variables 527 and control for others, and unlike in observational studies that apply guasi-experimental 528 statistical modelling approaches, observational studies with nominal estimates do not 529 standardize effort and do not explicitly account for simultaneous variability in potentially 530 informative predictors of a response (Cox et al. 2007; Luján and Todt 2021). Contrary to ACAP 531 (2023) criteria for identifying 'best practice' seabird bycatch mitigation methods, this is why it is important to validate that the efficacy of an intervention when used under controlled conditions 532 533 is of similar effectiveness when employed in real-world conditions through observational and 534 'pragmatic' studies (Khorsan and Crawford 2014; Pullin et al. 2021). Variability in tori line design 535 components that affect efficacy, including whether single or paired tori lines were employed 536 (Sato et al., 2013), may also have contributed to the relatively low efficacy of this mitigation method. The very small sample size for HookPods (9 records, see Figure S3) precludes any 537 538 meaningful assessment of relative efficacy. Furthermore, there may have been confounding 539 effects of other operational as well as environmental predictors of seabird bycatch risk that 540 effected the apparent relative efficacies of these 5 seabird bycatch mitigation methods - for 541 example, all HookPod records also had simultaneous use of night setting. Or, for example, crew 542 might employ a particular seabird bycatch mitigation method, such as a tori line or blue-dyed 543 bait, in areas and periods where high seabird bycatch rates are expected to occur, or in 544 response to observing high seabird interactions or high seabird density.

545 Due to the relationship between seabird density and bycatch risk, the apparent regional differences in relative seabird bycatch risk might be due to regional differences in the local 546 547 abundance of exposed species of seabirds that are susceptible to pelagic longline capture, and 548 thus regional differences in the magnitude of seabirds scavenging from longline vessels (Gilman 549 et al., 2005). The density of seabirds attending vessels might also affect their scavenging 550 behavior, where up to some threshold, the larger the local seabird abundance, the more intense 551 competitive scavenging behavior and risk of capture will be (Gilman et al., 2005). Regional 552 variability in the composition of seabird species complexes attending vessels might also partly 553 explain this finding. The effect of hierarchical competitiveness between seabird species and 554 between individuals within species, the diving behavior and capabilities of individual species in a complex attending vessels, as well as the occurrence and frequency of secondary interactions, 555 556 can affect seabird bycatch risk (Jimenez et al., 2012; Melvin et al., 2014). There were also likely 557 regional differences in a wide range of explanatory predictors of seabird bycatch risk that were 558 not able to be accounted for in our NMA, including operational characteristics of fishing methods and gear designs (including regional differences in the use of other seabird bycatch mitigation
 methods), environmental variables, seabird density and species composition of the assemblage
 attending vessels, and the spatiotemporal distribution of fishing effort.

562

563 4.3. Study Limitations

564 NMA is a powerful statistical procedure to assess multiple treatment effects within a coherent 565 modelling framework (Ades et al 2024). Nonetheless it is important to acknowledge the following 566 limitations when interpreting our network meta-regression model findings. We used the arm-567 based modelling approach as opposed to using treatment differences in a contrast-based 568 approach (Chu et al 2024) — the arm-based approach is far more flexible and it also supports 569 estimation of pairwise contrast effects (Ades et al 2024, Chu et al 2024). However, one ongoing 570 conceptual concern with arm-based approaches to NMA is that randomisation integrity within 571 trials or studies might not be preserved in the network and that this might bias estimated relative 572 treatment effects — known as the principle of concurrent control (Chu et al 2024) and related to 573 the pattern of missing treatment combinations within a study or trial since not all the trials 574 include all treatments as shown for example in Figure S2. This is an area of ongoing statistical 575 research.

576 Importantly, our study is based on study-specific aggregate or summarised data and not 577 set-specific data from each study. Consequently, all the study-specific covariates were also 578 marginal covariate summaries. This can lead to covariate imbalance across the 7 weight 579 designs and potentially increase the risk of bias in our model-based inference. This is also an 580 area of ongoing statistical research.

581 Moreover, while comprehensive, our study is nonetheless based on a small sample of 582 studies from a small number of fisheries. Sample size limitations prevented us from including various interaction terms in our network model such as ocean region-specific weighting design 583 584 effects, which might well be important. Furthermore, our NMA model could be mis-specified if 585 the branchline weighting designs were correlated with other potentially informative predictors of 586 seabird bycatch risk that were not available for inclusion in our study. Examples of such 587 potentially informative covariates include: seabird density attending vessels, species 588 composition of the exposed seabird complex attending the fishing vessels, seabird species 589 functional groups such as shallow or deep-diving species, environmental conditions, and other 590 operational characteristics of fishing methods and gear designs (Brothers et al., 1999; Jimenez 591 et al., 2014; Soriano-Redondo et al., 2016; Bi et al., 2020). Designs E and G had broad 592 definitions that could include a wide range of combinations of weight amounts and leader 593 lengths that could have large ranges of efficacies at reducing seabird bycatch risk — with larger 594 sample sizes these categories could be split into narrower design definitions.

596 4.4. Conclusions

595

597 Bycatch in pelagic longline fisheries is a major threat to some populations of pelagic seabirds. 598 particularly albatrosses and petrels (Anderson et al. 2011; Phillips et al., 2016). Managing 599 longline weighting designs is one approach for seabird bycatch mitigation that also supports 600 flexible compliance monitoring. Using a Bayesian multilevel network meta-regression modelling 601 approach, our study synthesized available evidence to assess the relative efficacies of 602 alternative branchline weighting designs, including those prescribed by RFMOs and ACAP. All 603 six assessed weighting designs were significantly more effective at reducing seabird bycatch 604 than a reference design of no weight within 5m of the hook, ranging from reductions of *ca.* 60% 605 (95% HDI: 7-96) to ca. 89% (95% HDI: 63-99) in the estimated seabird bycatch rates. There 606 was a >97% probability that all weighting designs significantly reduced seabird bycatch 607 compared to the reference design. But some RFMO and ACAP prescribed designs had 608 significantly different relative seabird catch rates. To augment seabird conservation, weighting 609 designs with relatively low or uncertain predicted efficacies could be assigned lower preferences

- 610 or eliminated. When combined with additional key inputs for bycatch management strategy
- 611 evaluation (Supplemental Material Section S8), these estimates of the relative efficacies of
- 612 alternative branchline weighting designs support the robust evaluation of alternative bycatch
- 613 management strategies for regional pelagic longline fisheries.
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- 615

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629 Supporting Information

- 630 This article includes online Supplemental Material.
- 631 632

633 **Conflict of Interest Statement**

634 The authors declare that they have no competing financial interests or personal relationships 635 that influenced the work reported in this article.

636 637

638 **Data Availability Statement**

639 Data from journal articles and grey literature materials used in this study are available in

640 Supplemental Material Table S2 and the original publications. Table S2 excludes records 641 derived from national observer programs included in the study for pelagic longline fisheries of

- New Zealand, Taiwan, U.S. Atlantic Ocean, U.S. Pacific Ocean shallow-set swordfish fishery, 642 643 and U.S. Pacific Ocean deep-set tuna fishery because these records are subject to 644 confidentiality restrictions that prevent the authors from making the observer data publicly available.
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883 Figure Captions

884

885 Figure 1. NETWORK GEOMETRY. Diagram showing the 7 (A-G) branchline weighting designs 886 at the nodes connected by 21 direct network interactions or edges between those nodes that 887 were sourced from 18 individual studies. This shows that our network is fully connected and 888 suitable for subsequent network meta-analysis. Node size is proportional to sample size with 889 connecting line (edge) thickness proportional to number of studies in our assembled dataset 890 that included summary data for comparison between any 2 specific designs (nodes). There are 891 three treatment (design) classes of: a reference design A (0 weight attached \leq 5m from hook), 892 designs B through F, which are prescribed by RFMOs or ACAP, and G is a non-prescribed 893 weighting design. Definitions of designs are in the Methods section 2.2 Branchline Weighting 894 Design Treatments.

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896 Figure 2. DESIGN-SPECIFIC CONDITIONAL EFFECTS. Summary of the predicted branchline 897 weighting design-specific conditional effects relative to the reference design A (no weight within 898 5m of the hook) based on the best-fit Bayesian network meta-regression model. Coloured 899 polygon shows the posterior density distribution summary (i.e., a wide and thin polygon 900 indicates low precision from a small sample size, and narrow and tall indicates there was high 901 precision and large sample size.), solid dots = median of the posterior density distribution, thick 902 horizontal line below each polygon shows the 80% highest posterior density interval (HDI) for 903 the density distribution while the thin horizontal line is the 95% HDI. Log rate (response) ratios < 904 0 = decreased seabird bycatch risk. Although a small fraction of 5 posterior density distributions 905 were > 0, (salmon shading) all the 6 contrasts with the reference design (A) were significantly < 906 0. 907

Figure 3. DESIGN-SPECIFIC RANK-ORDER. Posterior rank summaries for each of the
 branchline weighting designs (A-G) based on the best-fit network meta-regression model. Solid
 dot = posterior median rank, thin horizontal line = 95% credible interval summary of the
 neatoriar density distribution for that design aposition rank simulations.

911 posterior density distribution for that design-specific ranking simulations.

912

913 Figure 4. DESIGN-SPECIFIC CONSTRASTS. Summary of the predicted log rate ratios for all

21 branchline weighting design contrasts based on the best-fit network meta-regression model.

915 Solid dots = conditional median effect estimate, thick horizontal line = 80% credible interval

- while the thin horizontal line = 95% credible interval. Log rate ratio (response ratio) < 0 =
 decreased seabird bycatch risk.
- 918
- 919





923 Figure 1

Bayesian ML-NMR predicted branchline weighting effect

posterior density plots (with median and 80% & 95% HDI summaries)



posterior ranking of branchline weighting designs used to mitigate seabird bycatch in pelagic longline fisheries

(median with 95% credible interval summary)





relative effects for the 21 branchline weighting design contrasts

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